

1 **Geostatistical modelling of the spatial life history of post-larval deep-water hake**  
2 **(*Merluccius paradoxus*) in the Benguela Current Large Marine Ecosystem**

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## 29 **1. Introduction**

30 Deep-water hake (*Merluccius capensis*) is among the dominant demersal fish species in the  
31 Benguela Current Large Marine Ecosystem (BCLME), a productive upwelling system off the  
32 west coast of Southern Africa (5-37°S, 0-26°E). *M. paradoxus* inhabit the continental shelf slope  
33 from around 17°S in Angola/Namibia to about 27°E in South Africa (Figure 1) (Payne, 1995).

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35 *M. paradoxus* spawn mainly in areas of 200–650 m bottom depths between 34.5°S and  
36 36.5°S on the South African west coast. Spawning individuals has been found as far as 25°S in  
37 Namibia and 27°E off the South African south coast at depths ranging from 170 to 837 m (Jansen  
38 et al., 2015a). Spawning off the South African coast takes place throughout the year, with  
39 increased intensity around March and August–October (Jansen et al., 2015a).

40 Onshore, Offshore and alongshore ontogenetic migration has been indicated for *M.*  
41 *paradoxus* (Le Clus et al., 2005; Strømme et al., 2016). Based on catch rates by length class from  
42 South African surveys between 1990 and 2003, Le Clus et al (2005) concluded that *M.*  
43 *paradoxus* on the west and south coast was connected and likely belonged to the same stock.  
44 Strømme et al. (2016) extended this view to include Namibia where they found a large  
45 proportion of late juveniles, indicating a southern origin and a subsequent spawning migration to  
46 back to South Africa as the hakes matured. Juveniles initially migrate to shallower waters  
47 followed by a lifelong movement towards deeper waters (Botha, 1980; Burmeister, 2001; Gordoa  
48 and Duarte, 1991; Payne and Punt, 1995; Strømme et al., 2016). Large *M. paradoxus* has been  
49 found down to 1000 m (Burmeister, 2001; Mas-Riera, 1991).

50 The spatio-temporal spawning and migration patterns are in accordance with the most recent  
51 and comprehensive genetic study of *M. paradoxus* that found no spatial divergence and  
52 concluded in accordance with (Bloomer et al., 2009) that the population structure is one  
53 panmictic stock (Henriques et al., 2016). The authors found significant temporal divergence and  
54 suggested that this was from genetic chaotic patchiness. An earlier study had pointed towards  
55 multiple stocks (von der Heyden et al., 2010).

56 The area around the Lüderitz upwelling cell and Orange River Cone (LUCORC) region (25°-  
57 29°S) forms a natural barrier between the northern Benguela and southern Benguela (Agenbag

58 and Shannon, 1988; Lett et al., 2007; Rae, 2005). Therefore, and for practical/political simplicity,  
59 the national border between Namibia and South Africa (Orange River at 29 °S) is still regarded  
60 as the boundary dividing the *M. paradoxus* into two management stocks (Figure 1) (Burmeister,  
61 2005; Grant et al., 1988, 1987). The stock assessment and fisheries management advice has not  
62 yet adopted to the new understanding of *M. paradoxus* as one panmictic stock that migrate across  
63 the border between Namibia and South Africa. It is therefore unknown if the current fisheries  
64 practice for *M. paradoxus* in the BCLME is sustainable and optimal.

65 Recent developments in geostatistics and model fitting has facilitated analyses of large  
66 datasets from trawl surveys and let to new biological insight (Jansen et al., 2016, 2014;  
67 Kristensen et al., 2014).

68 In the present study we apply these methods to a large dataset of survey data from Namibia  
69 and South Africa to exhibit the spatial life history of post-larval *M. paradoxus*. This is the first  
70 analysis of *M. paradoxus* integrated both the South African, Namibian and Norwegian surveys.

71 We aim to test if the data supports the hypothesis of one panmictic stock and transboundary  
72 migration. Furthermore, we quantify the size specific migration for future usage in improved  
73 stock assessments.

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## 75 **2 Materials and methods**

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77 The data used for this study were the same as used by Jansen et al. (2016), except that data from  
78 2012 were not available for *M. paradoxus*. Modelling, post-processing and presentation of the  
79 results also followed the methods and design described in Jansen et al. (2016). The following  
80 description of materials and methods is therefore a modified copy from Jansen et al. (2016).

### 81 *2.1 Scientific trawl survey data*

82 The dataset used in this study comprised data from demersal trawl surveys conducted during  
83 the period 1998–2011 in the Benguela-Agulhas ecosystem between 17°S (northern border of  
84 Namibia) and 27°E on the South African south coast (Figure 1). All surveys sampled the  
85 demersal fish community on the continental shelf and upper shelf slope.

86 Namibian surveys used a Gisund Super demersal trawl towed by chartered commercial  
87 trawlers, all inter-calibrated with the *R/V Dr Fridtjof Nansen*, and the data were made available  
88 by the Ministry of Fisheries and Marine Resources (MFMR) in Namibia. Namibia surveys  
89 followed a systematic transect design with resampling of fixed transects with a semi-random  
90 distribution of stations along the transects. The South African surveys were conducted using  
91 either the “old” or the “new” configuration of a German 4-panel bottom trawl onboard *R/V*  
92 *Africana*. A few surveys were also conducted in South African waters using the Gisund trawl  
93 onboard *R/V Dr. Fridtjof Nansen*. South Africa surveys followed a random stratified sampling  
94 survey design. Details on the three types of gear, their operation and rigging are available in  
95 Axelsen and Johnsen (2015). Trawling time was approximately 30 minutes, and data from hauls  
96 shorter than 25 mins or longer than 35 were excluded from the analysis. Data from *R/V Africana*  
97 were made available by the Department of Agriculture Forestry and Fisheries (DAFF) in South  
98 Africa and data from *R/V Dr. Fr. Nansen* were provided by FAO/Norway (IMR). No trawl  
99 samples were available from Angolan waters due to species identification problems between *M.*  
100 *capensis*, *M. paradoxus* and *M. polli* (Benguela hake, which is largely caught in Angola).

101 For each trawl haul, the total catch was weighed and sorted to species level where possible.  
102 Large catches were subsampled. The weight of all *M. capensis* in the catch was recorded and  
103 total lengths of a subset of the individuals were measured in cm (rounded down). The length data  
104 recorded from the subsamples were subsequently raised to estimate the length composition in the  
105 total catch.

106 Data collected from nine hauls where winds in excess of 25 knots may have compromised  
107 the performance of the gear and hence the associated data points (Wieland et al., 2013) were  
108 excluded from the dataset.

109 The final survey dataset comprised 6 343 trawl hauls from 1998 to 2011, 57 % of which  
110 contained *M. paradoxus*. Catches (in numbers) ranged from 0 to 29 844 *M. paradoxus* per haul,  
111 with a mean catch of 867 in the non-zero hauls. The data set comprised a total of 3.1 million  
112 records of *M. paradoxus* lengths. Samples were available from the entire region except the  
113 shallow area (20 - 100 m) in central Namibia that is characterized by untrawlable bottom (Figure  
114 1). The samples were relatively equally distributed over the time-series (Fig. 2a). Most samples  
115 in Namibia and on the South African west coast were taken in January-February, while the South

116 African south coast was generally sampled in April-May (Figure 2b). Trawling was  
117 predominately done during the day (Figure 2c) at depths from 20 to 960 m (Figure 2d). The catch  
118 numbers from gears other than Gisund (i.e. “old” Africana and “new” Africana) were  
119 deterministically converted to the Gisund equivalent catch using the estimated length-specific  
120 gear effect from supplementary information 1.

121 The difference in the timing of the surveys in the various sub-regions in addition to the  
122 spawning phenology made it impossible to estimate the exact age of the hakes. *M. paradoxus*  
123 spawn throughout the year with increased intensity around March and August–October (Jansen  
124 et al., 2015a). The northern parts of the region (central Namibia) were surveyed in January to  
125 February (mean = 30 January). Consequently, the recruits in Namibia were primarily surveyed  
126 approximately 4.5 and 10.5 months after spawning. The South African west coast was also  
127 surveyed in January to February. Off the South African south coast surveys took place three  
128 months later (April-May, mean = 26 April). The young-of-the-year hake therefore entered the  
129 survey catches at various sizes and ages. For illustration purposes, mean age of the young-of-the-  
130 year hakes that entered the catches were set to 6 months.

## 131 2.2 The GeoPop model

132 Log Gaussian Cox (LGC) process modelling with correlations was used to describe  
133 abundance indices of *M. paradoxus* cohorts through space and time and along environmental  
134 gradients, observed using various gear types, as the hake recruited, grew and died. Similar  
135 models have previously proved their value for mackerel larvae and juveniles (Jansen et al.,  
136 2015b, 2012), cod (*Gadus morhua*) (Kristensen et al., 2013; Lewy and Kristensen, 2009),  
137 whiting (*Merlangius merlangus*) (Nielsen et al., 2014), and shallow-water hake (*Merluccius*  
138 *capensis*) (Jansen et al., 2016).

139 The first attempts to fit this type of model to the hake data from demersal surveys were  
140 unsuccessful because the large number of latent variables that needed to be estimated in this 4D-  
141 problem exceeded the limits of the physical processing power. The recent advent of the Template  
142 Model Builder (TMB; Kristensen et al., 2015) – a software library that can fit random effect  
143 models much faster than AD model Builder (Fournier et al., 2012) provided a solution to this  
144 obstacle. Physical processing power was furthermore increased by running the software on an  
145 r3.8xlarge instance under Amazon Elastic Compute Cloud (Amazon EC2). TMB was installed

146 on Linux on the r3.8xlarge platform consisting of 32 CPUs (2.5 GHz Intel Xeon E5-2670 v2,  
 147 AVX and turbo boost), 244 GiB RAM and 2 x 320 GB SSD storage  
 148 (<https://aws.amazon.com/ec2/instance-types/>). This setup facilitated the study of *M. capensis* in a  
 149 rich dataset by a complex model – a study that was impossible with ADMB or the model fitting  
 150 algorithms provided in the LGC-package used by Kristensen et al. (2013). The same setup was  
 151 used in the present study of *M. paradoxus*.

152 The model formulation and notation used for this study was identical to that described by  
 153 Kristensen et al. (2013). The response variable was catch in numbers from 47 size classes (2 cm  
 154 bins from 5 to 99 cm) in annual time steps in the period 1998 to 2011. The model consisted of  
 155 four main elements: The large scale spatio-temporal aspects ( $\eta(s, x, t)$ ), the local patchiness  
 156 ( $\eta_0(s, j)$ ), the population dynamics size-spectrum ( $\phi(s, t)$ ) and the gear selectivity ( $sel(s)$ ).  
 157 In the log-domain, the sum of these four components equals the total (latent) log intensity of the  
 158 poisson measurements indexed by haul-id  $j$ :

159  $\log(\lambda(s_j, x_j, t_j)) = \eta(s_j, x_j, t_j) + \eta_0(s_j, j) + \phi(s_j, t_j) + \log(sel(s_j))$  For the first three  
 160 elements the variance parameters are denoted by  $\sigma^2$ ,  $\sigma_N^2$  and  $\sigma_\varepsilon^2$ , respectively. Conditional on  
 161 the intensity  $\lambda$ , the counts (i.e. the number of *M. paradoxus* in each size class in a haul taken at a  
 162 particular point in space and time, with a particular gear) were assumed to be Poisson distributed  
 163 (Kristensen et al., 2013). This implied a resulting (unconditional) distribution with much higher  
 164 dispersion than the Poisson. This model structure is referred to as a log Gaussian Cox process  
 165 model and has been shown to be appropriate for count data from catches that are over-dispersed  
 166 and zero-inflated (with many zero values) (Jansen et al., 2015b; Kristensen et al., 2013). The  
 167 Poisson distribution allows for zero catches, while the log Gaussian randomness imply over-  
 168 dispersed catches (relative to Poisson), both allowing for very high counts and for many more  
 169 zero catches than would be found in a pure Poisson model.

170 The large scale spatio-temporal element ( $\eta(s, x, t)$ ) was a key feature of the model. It  
 171 models the time-varying heterogeneous spatial distribution of a size-structured population. The  
 172 statistical properties of this distribution were modelled in three parts: large-scale spatial patterns  
 173 (correlations between cells), temporal fluctuations (correlations between years) and size structure  
 174 (correlations between catches in different size classes). All correlations were assumed to decay  
 175 exponentially with spatial distance, time lag, and size difference, respectively, and assumed to be

176 multiplicative. The spatial correlation was assumed to be isotropic (direction independent). In  
 177 order to avoid correlation over land (e.g. the Cape region in South Africa), the spatial correlation  
 178 effect was modelled as a Gaussian Markov random field (cell-to-cell chains). The parameter  
 179 estimates for these correlations were expressed as decorrelation distance ( $H$ ), decorrelation time  
 180 ( $T$ ) and decorrelation size ( $L$ ), which were the distances in space, time or size where the  
 181 correlations had decayed to  $e^{-1}$  (explaining ca. 14% of the variance). Documentation of these  
 182 correlation structures were published in Kristensen *et al.* (2013). The spatio-temporal fields were  
 183 represented by annual time steps covering the 14 years of the study period (1998 – 2011) and the  
 184 spatial grid consisted of 278 cells each measuring  $50 \times 50$  km. The large scale spatio-temporal  
 185 model element therefore consisted of 182 924 random variables (47 size class  $\times$  14 years  $\times$  278  
 186 grid cells), which were assumed to follow a log Gaussian distribution that determined the mean  
 187 of the catch (in numbers).

188 The second main element in the model, which reflected local patchiness and fish' tendency to  
 189 aggregate with similar sized individuals on a local scale was the “nugget effect” ( $\eta_0(s, j)$ ). It is  
 190 a well-known phenomenon that catches of certain fish sizes tend to be overrepresented in trawl  
 191 hauls compared to the size distribution in the sampled population in the sea, likely due to the  
 192 size-structured nature of schools or shoals (Kristensen *et al.*, 2013). This local effect was  
 193 accounted for by modelling the remaining variation among the hauls (within cells) with a  
 194 Gaussian-distributed mean-zero term of variance  $\sigma_N^2$  and decorrelation range ( $L$ ) across sizes.

195 The third main element of the model was the stochastic population dynamics size spectrum  
 196 ( $\phi(s, t)$ ). It was governed by the McKendrick-von Foerster equation ( $\frac{\partial \phi}{\partial t} = -g \frac{\partial \phi}{\partial s} - z$ ) (Kot,  
 197 2001) with an added noise term ( $\mathcal{E}$ ) applied in every time step to make this model much more  
 198 flexible than the deterministic counterpart (Kristensen *et al.*, 2013). The constant growth and  
 199 mortality rates should therefore be interpreted as averages rates over time. The population  
 200 dynamics were discretized into 47 size classes and 32 time steps per year. Consequently, this was  
 201 represented by 21 056 random effects (47 size class  $\times$  14 years  $\times$  32 time steps).

202

203 Finally the catch was affected by the catchability of the gear ( $sel(s)$ ) and this effect was  
 204 implemented as:

205 
$$sel(s) = (1 + 3^{-(2/SR)*(s-\ell_{50}^{Gisund})})^{-1},$$

206 where  $sel(s)$  was the selection factor,  $SR$  was the selection range and  $\ell_{50}^{Gisund}$  was the fish length  
 207 ( $s$ ) in cm at half selection for the Gisund gear.

208 The parameters in the model were estimated using the maximum likelihood principle based  
 209 on the Laplace approximation and thus the estimation follows the principles of Kristensen et al.  
 210 (2013).

211 The fitted model was then used to calculate annual estimates of the abundance index of hake  
 212 in each size class in each 50 x 50 km cell. These spatio-temporal distribution patterns were also  
 213 transformed into a coastline-oriented coordinate system, to reveal any along-shore migration  
 214 patterns. This was done by projecting the estimated abundances in the Cartesian coordinate  
 215 system onto an axis that followed the coastline as a straight line from the Namibia-Angola border  
 216 (11°E, 17°S) in the North, via Cape Point (18.5°E, 34°S) to east of Port Elizabeth on the South-  
 217 African south coast (27.5°E, 34°S) (see Fig. 1). Multiannual average distributions were  
 218 calculated as the unweighted average distribution of multiple years.

219 For explanatory purposes, the length-based model outputs were presented by both length and  
 220 absolute age. The length-to-age conversions were done by using the age readings by DAFF (D.  
 221 Durholtz, DAFF, unpublished data). This was done by calculating the average age by length  
 222 from the Von Bertalanffy equations fitted to age readings of male and female *M. capensis*,  
 223 respectively (plot provided in results section). Maps and bar plots were constructed to show  
 224 mean distributions of *M. capensis* of length groups corresponding (+/- 1.5 cm) to the ages in one  
 225 year steps from 0.5, the mean size and age when a new cohort first appears in the surveys, to 9.5  
 226 years.

227 Depth data were downloaded as gridded (0.07° lat x 0.07° lon) averages from the NOAA  
 228 ETOPO1 database (Amante and Eakins, 2009) using the “marmap” package (Pante et al., 2015).  
 229 Mean depth was calculated for each cell and isobaths were produced for the maps. Finally,  
 230 abundance in cells with a mean depth exceeding the deepest trawl haul (960 m) was set to 0. The  
 231 deepest observed *M. paradoxus* catch was at 934 m.

232



### 233 3. Results

234 The parameter estimates (and associated standard errors) of the model fitted to the catch data  
235 are given in Table 1.

236 Hake densities were found to be correlated in space and size with a spatial de-correlation  
237 distance ( $H$ ) of 481 km and a length de-correlation difference ( $L$ ) of 79 cm. These patterns were  
238 found to be very stable from year to year with a temporal de-correlation period ( $T$ ) spanning 30.1  
239 years ( $CV = 7\%$ ) and could therefore be examined as averages over the entire time series.

240 Estimated variance parameters revealed a roughly equal contribution from spatial large-scale  
241 patterns ( $\sigma^2 = 20.9$ ) and the nugget effect ( $\sigma_N^2 = 12.4$ ). In comparison, the contribution from  
242 population dynamics was much smaller ( $\sigma_\varepsilon^2 = 0.03$ ). The variance in abundance of a given size  
243 class at a single point in time and space was therefore completely dominated by large scale  
244 spatial effects and local effects rather than demographic effects.

245 The total mortality rate of *M. paradoxus* was estimated to  $0.33 \text{ year}^{-1}$  ( $CV = 67\%$ ).  
246 However, this varied through life and between areas. The mortality increased substantially when  
247 the fish exceeded approximately 55 cm (Figure 3). The growth rate was estimated to be  $8.3 \text{ cm}$   
248  $\text{year}^{-1}$  ( $CV = <0.001\%$ ). Despite the very weak population effect, these estimates (mortality and  
249 growth) resembled previously reported rates (Figure 3 and 4). However, the cohort signals were  
250 too weak to be used for tracking of single cohorts and their effects on spatio-temporal  
251 distribution fields were negligible.

252 *M. capensis* was not fully recruited to the trawl survey in their first year (Figure 3). This was  
253 evident from the estimated selection of the reference gear “Gisund” ( $\ell_{50}^{Gisund} = 8.4 \text{ cm}$ ,  $SR^{Gisund}$   
254  $= 1.4 \text{ cm}$ ) and the increase in abundance from 0.3 to 1.3 years (Figure 3).

#### 255 3.1 Spatial patterns (distribution, migration, population structure)

256 The spatial distribution of the catch rate of *M. paradoxus* was estimated and illustrated to  
257 infer age and size-related migration patterns. This was done for length groups corresponding to  
258 the ages from 0.3, the approximate age when a new cohort first appears in the surveys, to 9.3  
259 years.

260 To examine the alongshore migration patterns, we projected the length-age-specific spatial  
261 distributions onto an axis consisting of two straight lines, one from the Namibia-Angola border

262 (Kunene River mouth) in the North, via Cape Point to Port Elizabeth in the south-east. The  
263 alongshore projections (Figure 5 and 6) and distribution maps (Figure 7 and 8) indicated  
264 horizontal movements throughout the life span of *M. paradoxus*. The initial distributions of the  
265 approximately 0.3- year-olds (recruits) and the 1.3-years olds (juveniles) indicated one main  
266 nursery area (Figure 6 and 7). During the first 7 years of the life of *M. paradoxus*, the  
267 distribution shifted gradually from year to year, indicating considerable alongshore migrations.  
268 In the north, considerable quantities of *M. paradoxus* appeared to move northwards from the area  
269 around Orange River - their most likely origin. In lesser number, *M. paradoxus* moved eastwards  
270 along the south coast of South Africa. *M. paradoxus* were most widespread at the age of  
271 approximately 4.3 years (51 cm) where almost 100 % were mature (50 % mature at 42 cm (Singh  
272 et al., 2011)). Later, the distributions progressively contracted in the vicinity of the nursery area  
273 (Figure 5, 6 and 8). The movements north of Orange River (the border between Namibia and  
274 South Africa) and east of Agulhas thus indicated natal homing as the most parsimonious  
275 explanation. Consequently, transboundary movements most likely occurred.

276 *M. paradoxus* appeared initially to move to shallower waters at around 80 m depth, after  
277 which they gradually moved deeper from approximately 0.5 to 5 years of age. From the age of 5  
278 they were found from 150 to 935 m, but mainly between 350 and 650 m, while moving into  
279 slightly shallower waters (Figure 9).

#### 280 **4. Discussion**

281 The present analysis was based on the hitherto largest database of *M. paradoxus* survey data,  
282 for the first time including surveys from both the Namibian, Norwegian and South African  
283 fisheries research institutes. The quantitative geostatistical modelling exposed spatial migration  
284 patterns largely in line with Strømme et al. (2016) that were based on visual inspection of  
285 presence/absence outline maps based. Like Strømme et al. (2016) and Le Clus et al. (2005) we  
286 assumed that the spatial shifts in distribution over the life span of post-larval *M. paradoxus* were  
287 primarily reflecting migration and to a lesser extent mortality, however, the effects of spatial  
288 variation in mortality remains to be studied.

289 *4.1 Spatial patterns (distribution, migration, population structure)*

290 The results indicated one primary recruitment/nursery area on the west coast of South Africa  
291 and a secondary low-production area around Port Elizabeth on the South Coast. Juveniles  
292 initially migrated away from the main recruitment area, followed by natal homing by larger  
293 individuals. This pattern was highly consistent through the studies time series as indicated by the  
294 very long decorrelation time.

295 The return migration from Namibia appeared to take place when the hakes exceeded 50 cm.  
296 At this size almost 100 % would be mature if they matured at the same size as in South Africa  
297 Singh et al (2011), which is in contrast to the very few findings of spawning *M. paradoxus* in  
298 Namibian waters (Jansen et al., 2015a). Size of maturation therefore needs to be estimated for the  
299 part of the stock that migrates into Namibian waters.

300 At the South coast of South Africa *M. paradoxus* appeared to move slightly westwards and  
301 then stay after reaching the size of maturation. This corresponds to the spawning reported in this  
302 area by Jansen et al. (2015a). However, very few recruits were observed, indicating that either  
303 they were outside the surveys area, transported by the currents to the west coast before they were  
304 caught in the survey, underestimated (unaccounted gear avoidance due to factors such as vertical  
305 distribution or gear), or the reproduction in this area was less successful than on the west coast.

306 The spatial life history patterns thus conform to the concept of one main population unit  
307 (stock), but points to an additional smaller component on the eastern part of South Africa's south  
308 coast. The level of interconnection (mixing/straying) between these components (stocks) is  
309 presently unknown, but appears to exceed the threshold where genetic differentiation occurs in  
310 the mitochondrial DNA (Henriques et al., 2016). Furthermore, the rarity of small juveniles in the  
311 eastern survey catches is puzzling and calls for field work using other techniques than the  
312 standard survey.

313 *M. paradoxus* reached its deepest distribution at around 50-55 cm length (Figure 9), where it  
314 was abundant at depths deeper than 600 m. These depths were not completely covered by the  
315 surveys. This corresponds to an increased mortality at the same sizes (Figure 3), which could be  
316 explained by fish moving out of the survey coverage.

317

318 *4.2 Population dynamics and nugget effects*

319 The effect of the population dynamics in the model was minuscule, this could be a  
320 consequence of the year-round spawning and recruitment of *M. paradoxus* combined with the  
321 difference in timing of the South African South coast and West coast surveys. Despite this,  
322 growth and mortality rates appeared to be relatively well estimated as indicated by the low  
323 standard deviations of the estimates and the comparison with values reported from age readings  
324 and stock assessments. A plausible explanation for this result, which is consistent with what was  
325 found for *M. capensis* (Jansen et al., 2016), is that the variability in the size-structured spatial  
326 patterns is so relatively large, that the contribution from the population dynamics is not required  
327 to explain patterns in data.

328

329 The data and model fit for *M. paradoxus* did not indicate any large spatial differences in  
330 small scale variability (nugget effect) that complicated the interpretation of the model fit like for  
331 *M. capensis* (Jansen et al., 2016). This is evident from supplementary information 2 that is made  
332 for comparison with figure 5 and 6 in Jansen et al. (2016).

333

334 **To be added:** 1. Discuss gear effects. 2. Test sensitivity to the gear effect estimates. How much  
335 would the results change if the gear effects changed (Supplementary information 3).

336

#### 337 *4.2 Model performance and future developments*

338 The GeoPop model has been made possible by Template Model Builder's computational  
339 approach to general statistical problems with many unobserved random variables, as well as the  
340 availability of memory and processing power through cloud computing. GeoPop represents a  
341 major improvement of earlier approaches, specifically by including both population dynamics,  
342 large-scale spatial patterns, and small-scale size-structured clustering. Since GeoPop lies at the  
343 boundary of what is currently computationally feasible, a number of compromises had to be  
344 made with respect to the ecological fidelity of the model, and it is foreseeable that future  
345 developments in algorithms and computational resources will allow further refinement of the  
346 model assumptions. Here, we list three candidates for such refinements:

- 347 1. Currently, the correlation structure in the spatial fields is assumed to be isotropic, i.e.  
348 independent of direction. It is plausible that correlations between different locations  
349 depend not only on distance, but also on differences in depth or other key habitat  
350 parameters. Some extensions are relatively straightforward and could for example be  
351 achieved by changing coordinates from latitude and longitude to alongshore distance and  
352 depth. However, this solution would have other flaws, as it would e.g. regard an off-shore  
353 bank as an extension of the near-coast space. A more general and robust structure could  
354 be obtained by modelling the fluxes in space that drive the redistribution of the stock  
355 between time points; this could include diffusive fluxes which would be anisotropic, as  
356 well as advective fluxes which indicate preferred migrations. While such an extension is  
357 appealing from the point of view of first principles in spatial ecology, by explicitly  
358 modelling migrations, it would lead to correlation structures that make computations  
359 much more demanding. In general, inference in spatiotemporal dynamics, including  
360 model selection, is a computationally intensive task in which the trade-off between  
361 fidelity and feasibility must constantly be reassessed as computational abilities expand.
- 362 2. The growth dynamics and the mortality patterns are grossly simplified, in that both are  
363 assumed constant. A major effort would be required to allow changes in growth rates,  
364 explained by current size and possibly also environmental covariates, as well as varying  
365 mortality, most importantly described by spatiotemporal patterns in fishing effort.  
366 Ultimately, the inclusion of such patterns could lead to a spatiotemporal size-based  
367 assessment model. The effort required for this development would be very substantial,  
368 but single steps in this direction could be pursued.
- 369 3. It could be envisioned to include different stock components, differing in timing and  
370 location of spawning. This would facilitate the tracking of individual cohorts.

371

## 372 **5. Conclusion**

373 The perception of the migration and population patterns derived from the first geostatistical  
374 modelling of data from all surveys in the region corresponds largely to the hypothesis and data

375 plots presented by Le Clus et al. (2005) and Strømme et al. (2016), but differ from the current  
376 assessment practices in Namibia and South Africa, where two stocks are assumed to be separated  
377 by the national border.

378

379 **Acknowledgements**

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482

483

484 Table 1. Model parameter estimates and standard errors. Parameters with unit “1” are  
 485 dimensionless.

486

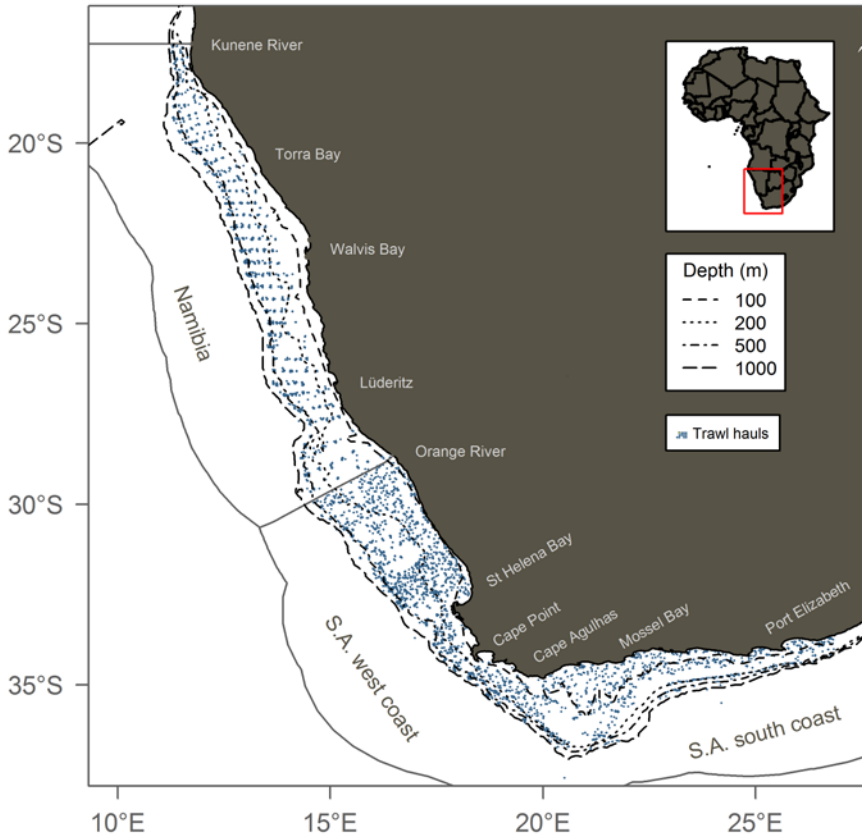
Symbol	Description	Unit	Estimate	Standard error	CV (%)
$\ell_{50}^{Gisund}$	Fish size at half selection (Gisund)	cm	8.372	0.882	11
$SR^{Gisund}$	Selection range (Gisund)	cm	1.414	0.308	22
$g$	Growth rate	cm/year	8.320	<0.001	<0.001
$z$	Total mortality	1/year	0.325	0.218	67
$\sigma_{\varepsilon}^2$	Variance of population dynamics	1	0.033	0.006	18
$T$	Decorrelation time	year	30.088	2.220	7
$L$	Decorrelation size	cm	79.069	1.894	2
$H$	Spatial decorrelation distance	km	480.606	31.087	6
$\sigma^2$	Variance of the space-time-size correlations	1	20.934	1.265	6
$\sigma_N^2$	Variance of the nugget effect	1	12.350	0.283	2

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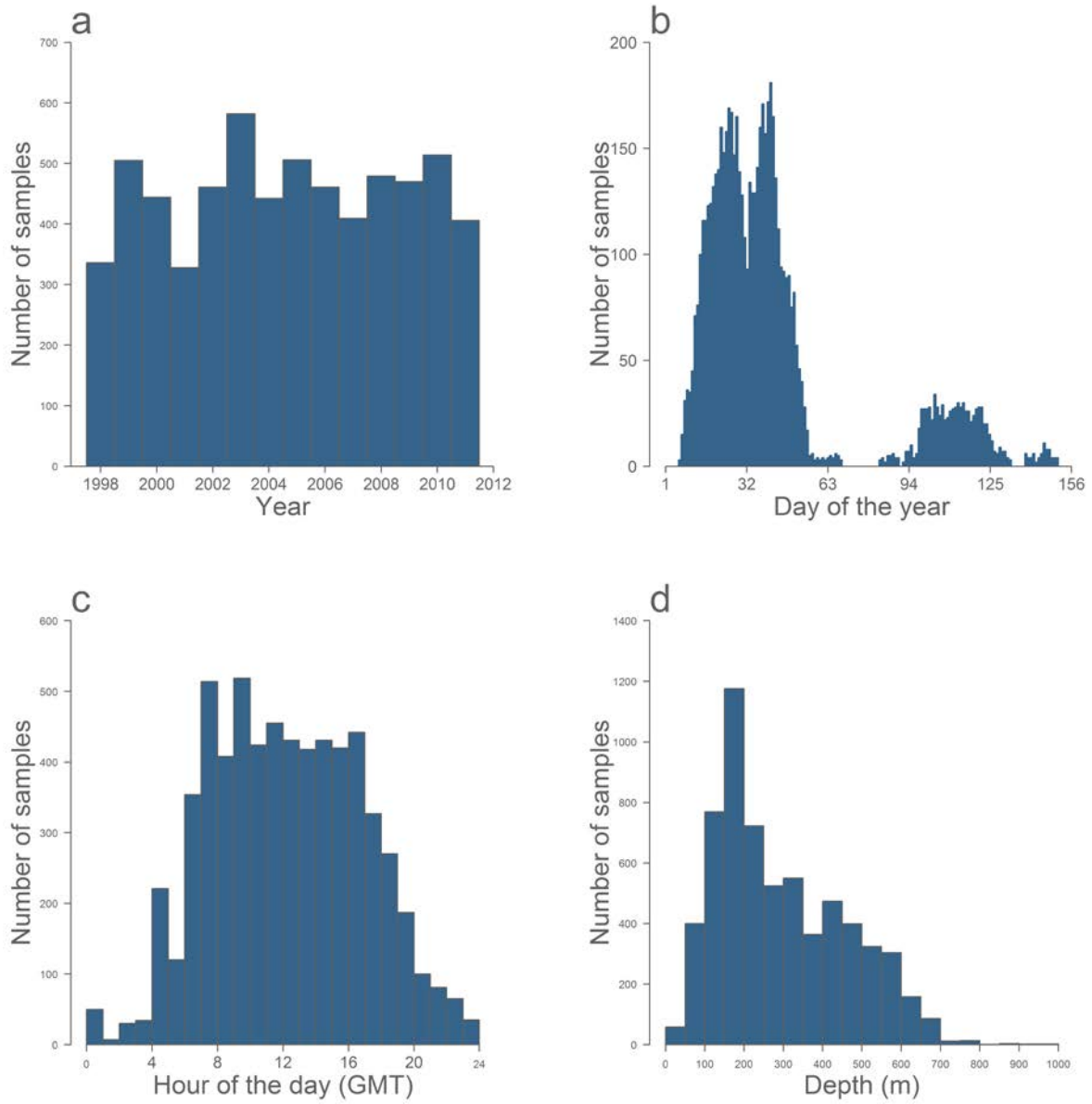
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490 **Figures**  
 491



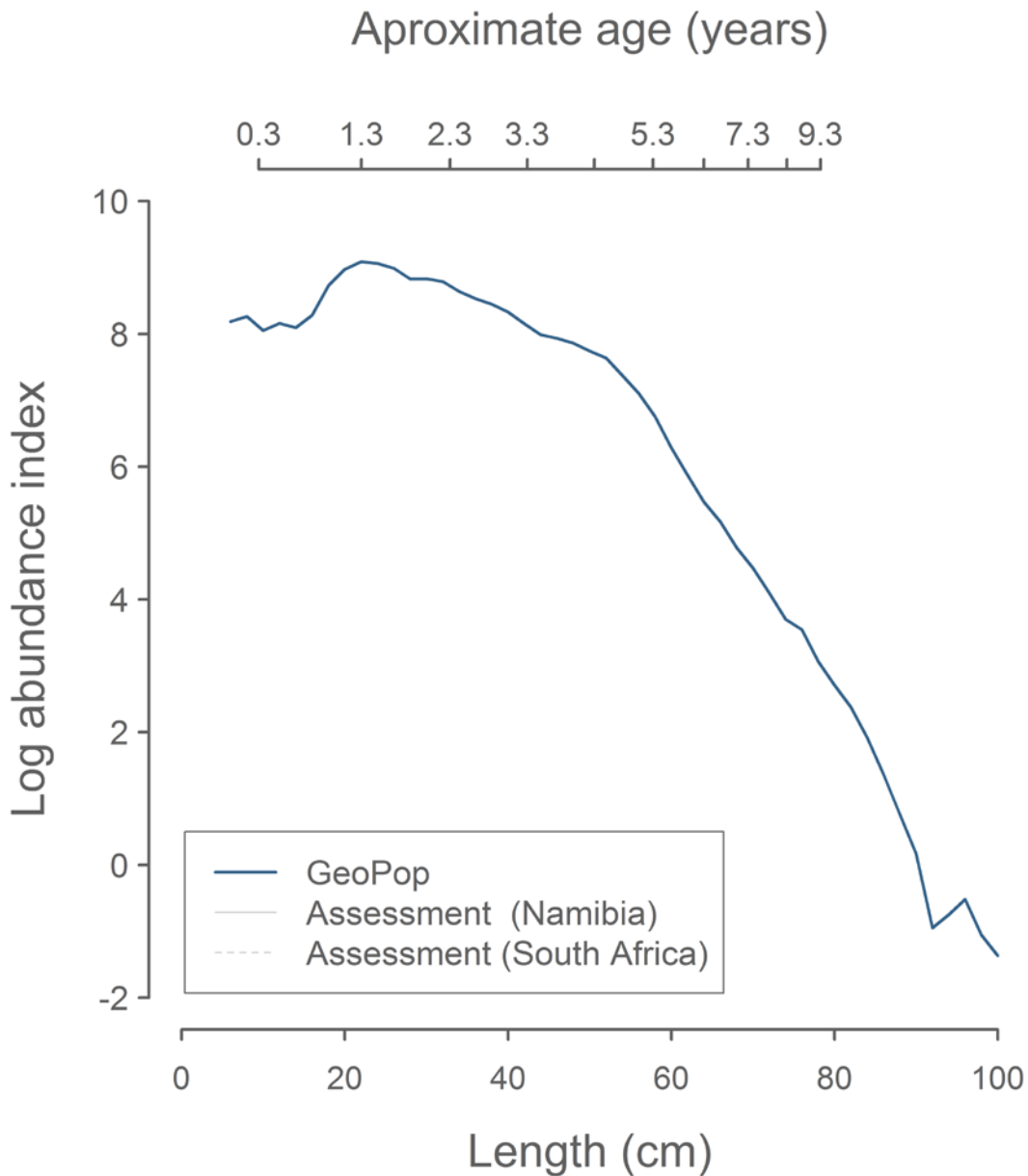
492  
 493 Figure 1. Map of study area with sample locations (dots), isobaths and place names referred to in  
 494 the text.  
 495



496

497 Figure 2. Demersal trawl survey samples (trawl hauls) by year (a), ordinal day (b), hour of the  
 498 day (c) and depth (d).

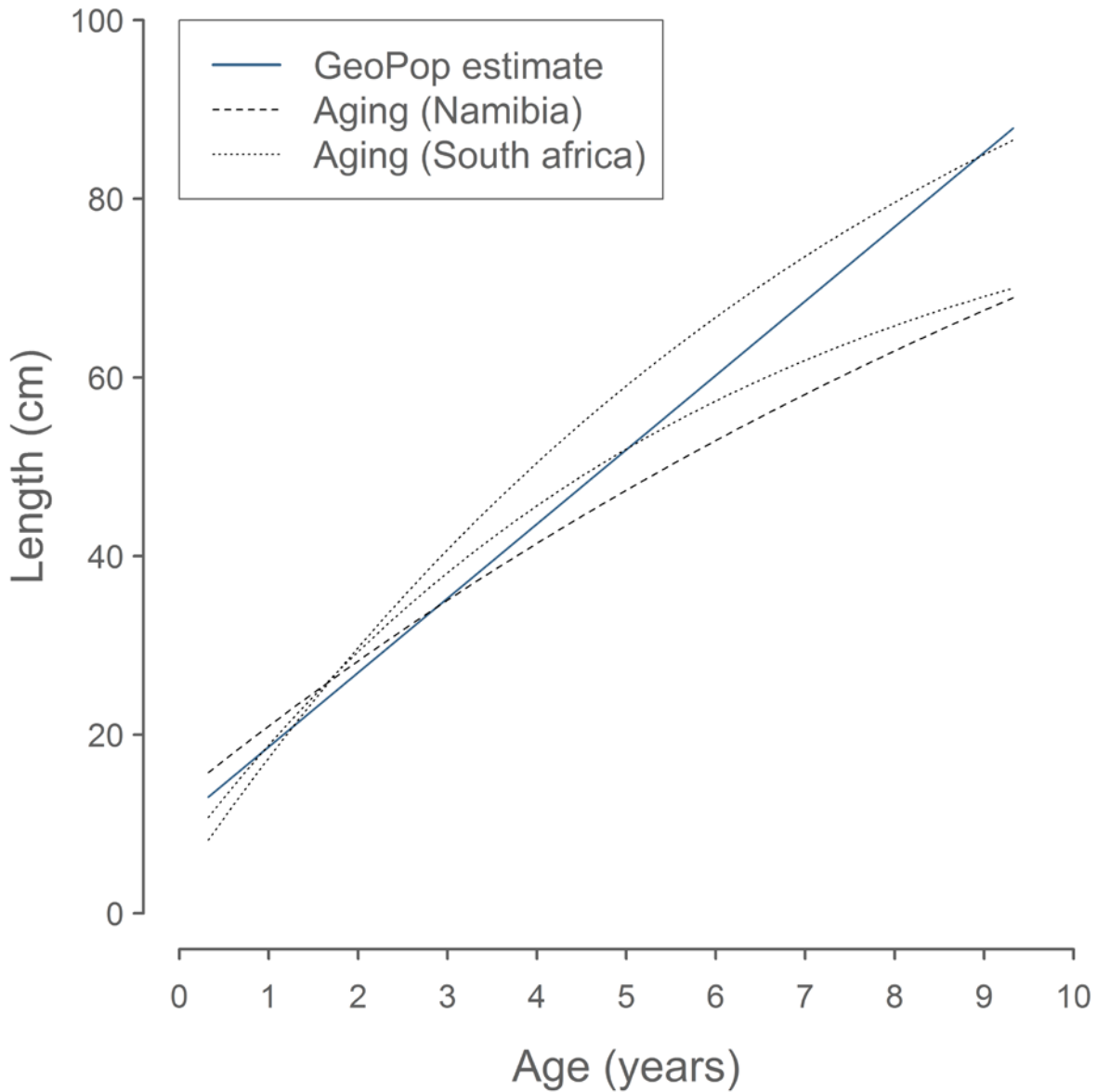
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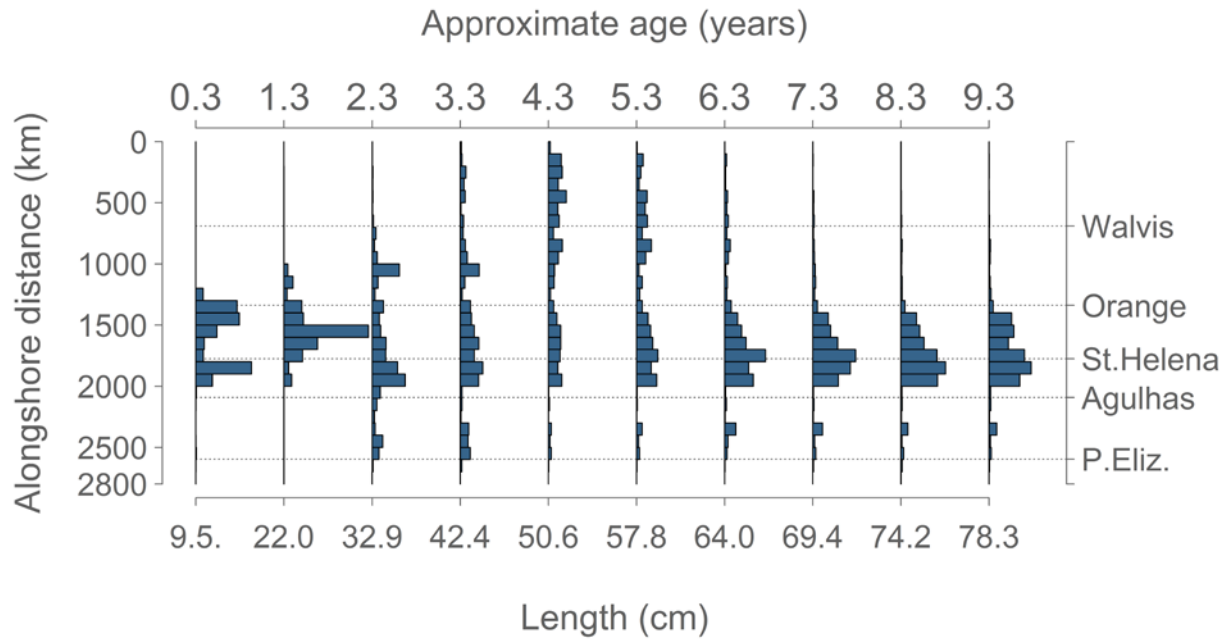
500

501 Figure 3. Log abundance index of deep-water hake (*M. paradoxus*) by length and approximate  
 502 age (blue line). Inclination equals the total mortality. The mortality estimated in the  
 503 Namibian and South African stock assessments are indicated by grey lines [To be added]  
 504 (mean Z in South Africa from 1998 to 2012 from Rademeyer (2012) and in Namibia from  
 505 1998 to 2011 from Kathena et al. (2015)).

506



507  
 508 Figure 4. Comparison of growth rate estimates for deep-water hake (*M. paradoxus*). Estimate  
 509 from the GeoPop model ( $8.4 \text{ cm year}^{-1}$  from first surveyed at an age of approximately 0.3  
 510 years corresponding to 9.5 cm according to South African age readings, blue straight line),  
 511 Von Bertalanffy equations fitted to South African age readings (females: upper dotted line;  
 512 males: lower dotted line), Namibian age reading (dashed line).  
 513



514

515 Figure 5. Alongshore distribution of number of deep-water hake (*M. paradoxus*) by length and

516 approximate age estimated by GeoPop model after removal of local noise (nugget effect).

517 The distributions are standardized for each age, so the areas of the bars are the same for all

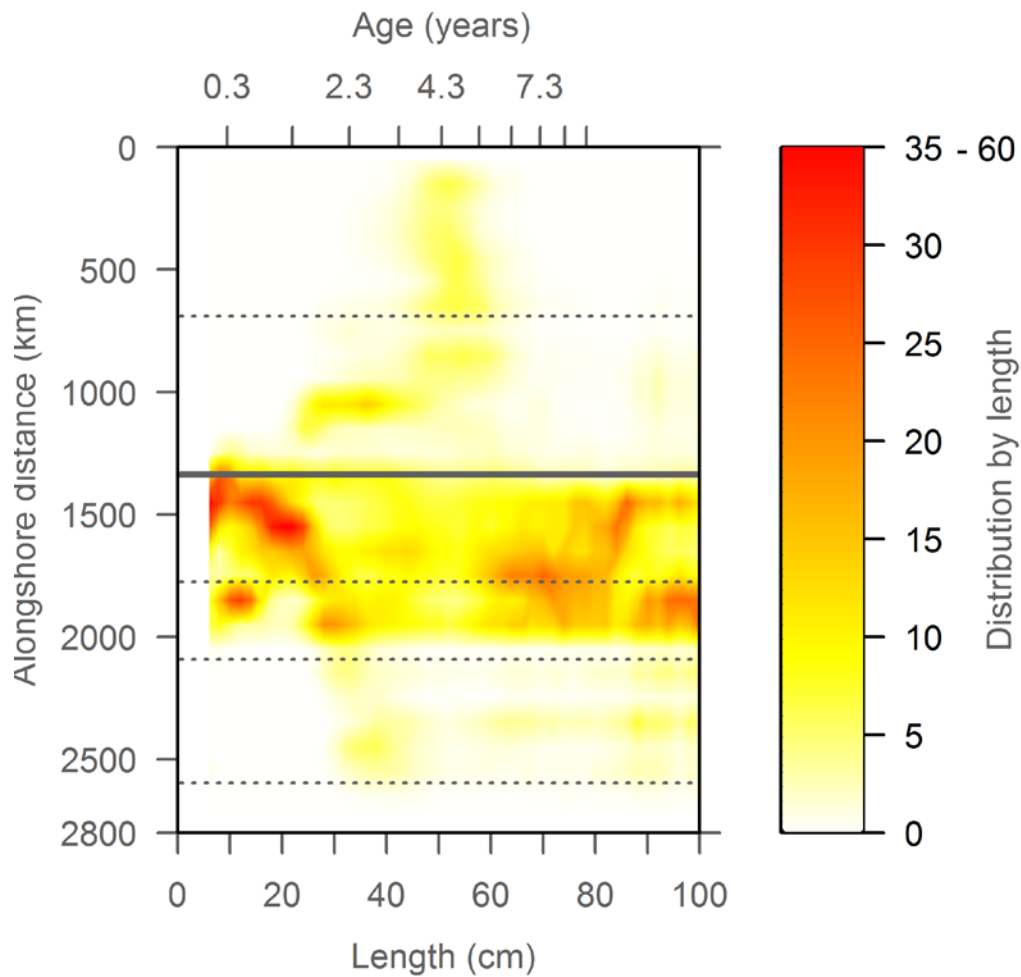
518 age distributions. The spatial distributions were projected onto a curvilinear axis following

519 the coastline from the Namibia-Angola border (Kunene River) in the North to Port Elizabeth

520 in the south-east.

521





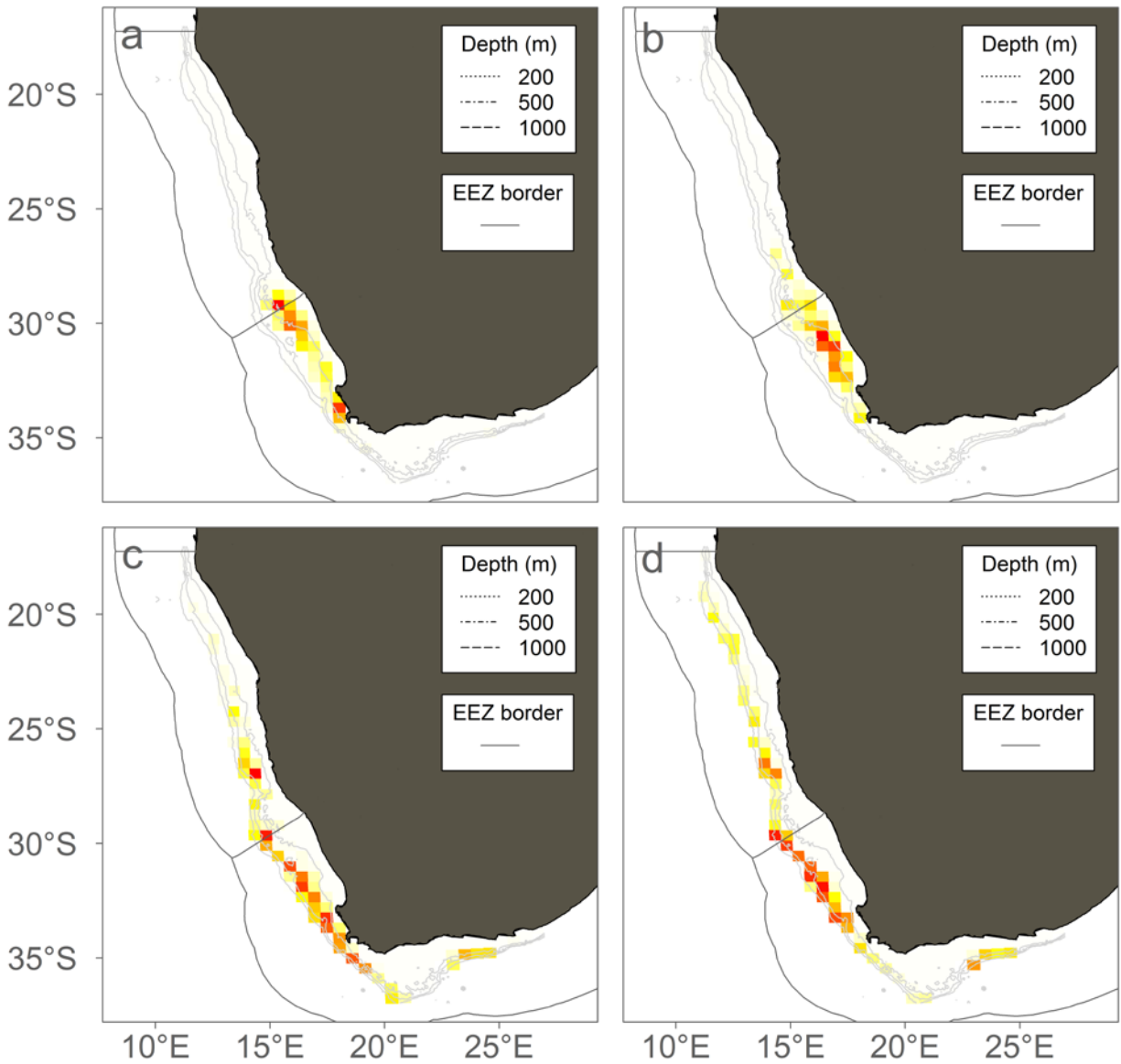
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523 Figure 6. Alongshore distribution of number of deep-water hake (*M. paradoxus*) by length and  
 524 approximate age estimated by GeoPop model after removal of local noise (nugget effect).

525 The distributions are standardized for each 2-cm length class. The spatial distributions were  
 526 projected onto a curvilinear axis following the coastline from the Namibia-Angola border

527 (Kunene River) in the North to Port Elizabeth in the south-east.

528



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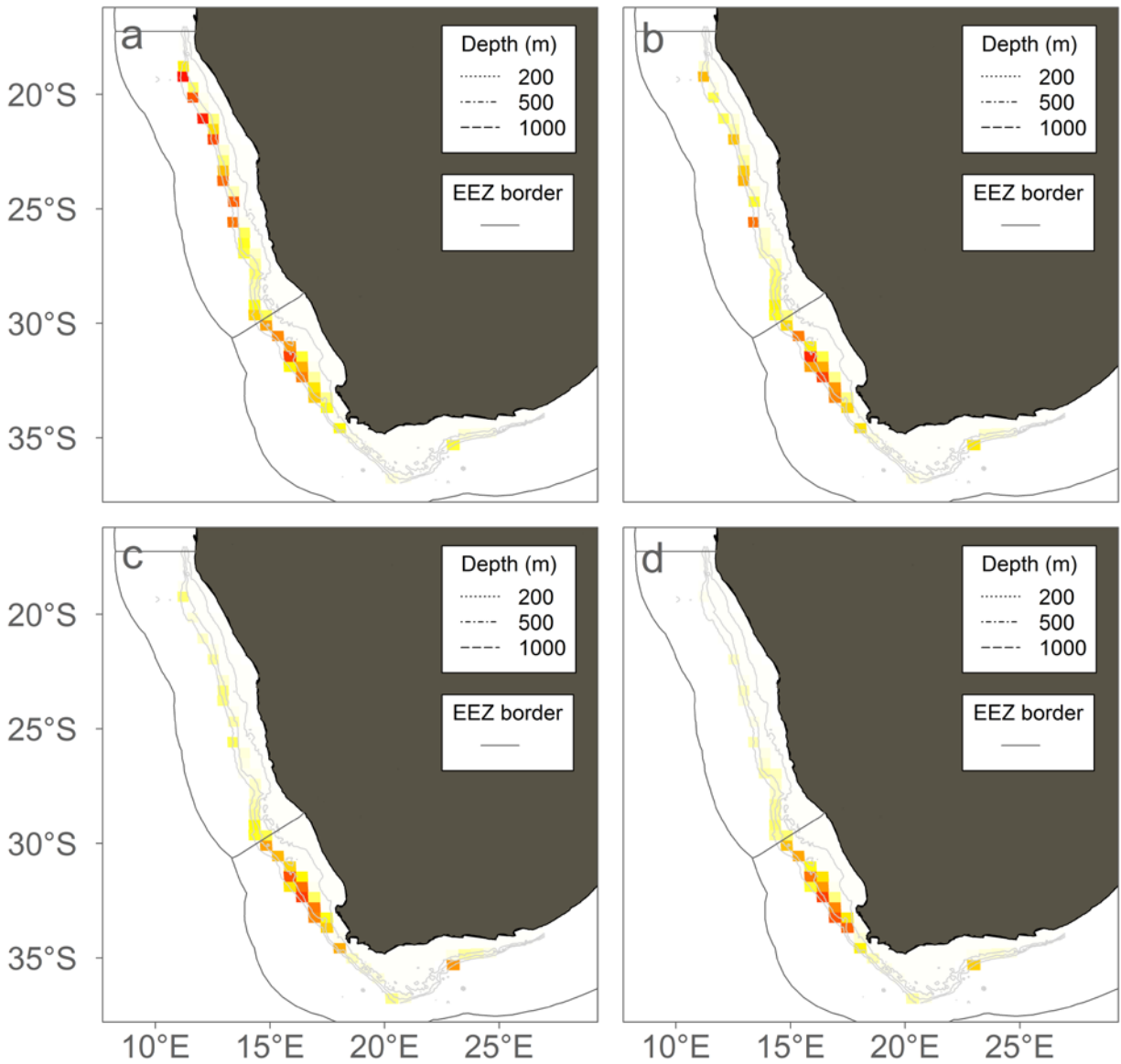
530 Figure 7. Distribution maps of deep-water hake (*M. paradoxus*) at the age of approximately 0.3

531 years (a), 1.3 years (b), 2.3 years(c) and 3.3 years (d). The distributions are illustrated as

532 cumulative fractions, e.g. the sum of all areas with the color corresponding to 40 %

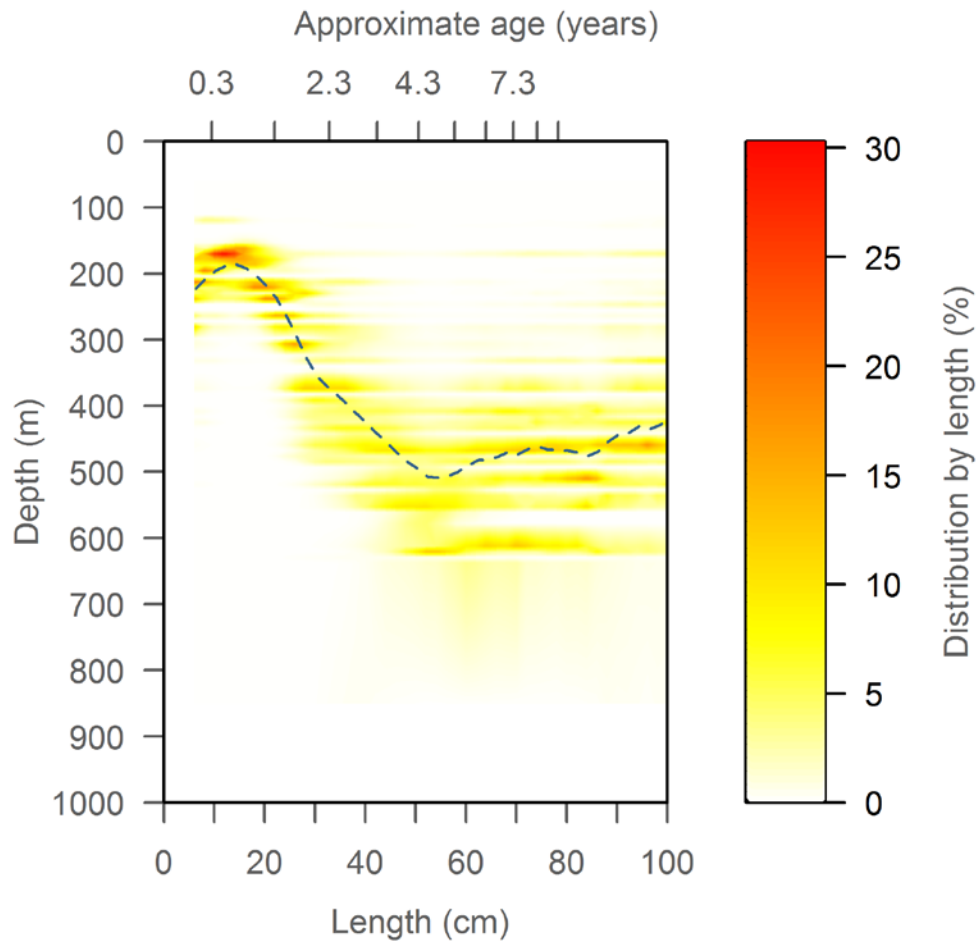
533 represents 40 % of the total.

534



535  
 536 Figure 8. Distribution maps of deep-water hake (*M. paradoxus*) at the age of approximately 4.3  
 537 years (a), 5.3 years (b), 6.3 years(c) and 7.3 years (d). The distributions are illustrated as  
 538 cumulative fractions, e.g. the sum of all areas with the color corresponding to 40 %  
 539 represents 40 % of the total.

540



541

542 Figure 9. Depth distribution of deep-water hake (*M. paradoxus*) by length and approximate age.

543 Blue dashed lines indicate the weighted mean depth.

544

545 **Supplementary information**546 Supplementary information 1. The gear effect on catch rates of *M. paradoxus*.

547

548 Supplementary information 2. No large scale spatial bias in nugget effect.

549

550 Supplementary information 3. Sensitivity test for the gear effect on the catch rates of *M.*551 *paradoxus* [Not done yet].

552

553

554

### Suppl. info. 1: The gear effect on catch rates of *M. paradoxus*

As detailed in Materials and Methods, the data base for the present study includes catch rates with different gears. The study must take the difference in the size selectivity and efficiency of the different gears into account, to avoid spurious patterns and bias in the estimated spatial distributions of the stock. Here, we describe how the catches from the R/V Africana are converted to equivalent catches that we can assume would have been obtained with Gisund. We refer to this as “gear intercalibration”.

We constructed a statistical method for intercalibration, i.e. determining the relative selectivity of two gear types, based on data from paired trawl hauls. The model estimates the size spectrum of the underlying population at each station, size-structured clustering of fish at small temporal and spatial scales, in addition to the relative selectivity of the two gears in each length class. The statistical assumption is Poisson distributed catches conditional on log-Gaussian variables that describe the expected catches, which allows for overdispersion and correlation between catch counts in neighboring size classes.

#### SI 1.1: Statistical model

The intercalibration model is a statistical model which explains the size composition of the catch in survey trawl hauls. The model is a non-linear mixed effect model, in which we do inference using numerical maximum likelihood estimation, employing the Laplace approximation to integrate out random effects.

The observed quantities are count data,  $N_{ijk}$ , which represents number of individuals caught at station  $i = 1, \dots, n_s$ , with gear  $j = 1, 2$ , and in length group  $k = 1, \dots, n_l$ . Here, the length groups are 2 cm length classes starting at 10 cm.

We assume that these catches are Poisson distributed, conditional on the swept area  $A_{ij}$  and three sets of random variables, which all depend on the size class  $k$ : First, the local background size spectrum  $\Phi_{ik}$ , which is specific to the station, second, haul-specific fluctuations  $R_{ijk}$  in the size spectrum, and third, the relative selectivity  $S_{jk}$  which is specific to the gear. More specifically,  $\Phi_{ik}$  represents the size composition of the fish at station  $i$ , as would be observed with a hypothetical gear with “typical” size selectivity, so that  $\exp(\Phi_{ik})$  is the expected number of fish caught in size group  $k$  at station  $i$  with a hypothetical gear which lies in between the two gears  $j = 1$  and  $j = 2$ .

Next, the haul-specific fluctuations  $R_{ijk}$  are akin to the “nugget effect” in spatial statistics, and represents small-scale clustering of fish. This is particular to both stations and gear, since the paired hauls are done at slightly different locations and times, and therefore these clusters have moved or regrouped between hauls at the same station.

Finally, the selectivity  $S_{jk}$  is the main object of interest, and represents the selectivity of gear  $j$  in size group  $k$ . Since we do not know the actual size

distribution of the stock, we cannot estimate the absolute selectivity, but only the relative selectivity between the two gears. This corresponds to enforcing  $S_{1k} = -S_{2k}$

Given these random variables  $\Phi$ ,  $R$ ,  $S$ , we assume that count data is Poisson distributed:

$$N_{ijk} | \Phi, R, S \sim \text{Poisson}(A_{ij} \cdot \exp(\Phi_{ik} + S_{jk} + R_{ijk}))$$

The swept area  $A_{ij}$  is an input to the model. The unobserved random variables,  $\Phi$ ,  $R$  and  $S$ , are given prior distributions: The size spectrum at each station, i.e.  $\Phi_{ik}$ , is considered a random walk over size groups:

$$\Delta \Phi_{ik} \sim N(0, \sigma_{\Phi}^2) \text{ for } k = 1, \dots, n_l - n_{\Phi} \quad .$$

Here,  $\Delta$  is the difference operator. This enforces continuity in the size spectrum; the most probable spectrum is flat. To ensure that the spectrum is a well defined stochastic process, we complement this with initial conditions

$$\Phi_{ik} \sim N(0, \sigma_1^2) \text{ for } k = 1 \quad .$$

Here, the variance  $\sigma_1^2$  is fixed at a “large” value 10. In contrast, the parameter  $\sigma_{\Phi}^2$  is estimated. We assume independence between stations, i.e. we do not attempt to model any large-scale spatiotemporal structure of the population. We note that this is the main difference between this model and the GeoPop model, where emphasis is exactly on this spatiotemporal structure.

The residual or “nugget effect”  $R_{ijk}$  models size-structured clustering of the fish at small spacial and temporal scales. Thus, this effect is independent between hauls, even those taken at same station  $i$  but with different gear  $j$ . For a given haul, i.e. for given station  $i$  and gear  $j$ , the nugget effect is a mean 0 first order autoregressive process of size, with a variance  $\sigma_N^2$  and correlation coefficient  $\phi$  which is estimated.

The relative selectivity  $S_{jk}$ , which we aim to estimate, is modeled as a random walk in size:

$$\Delta S_{jk} \sim N(0, \sigma_S^2) \text{ for } k = 1, \dots, n_l - n_S$$

We assume infinite variance on the first size group,  $S_{j1}$ , i.e. only the increments in the selectivity process enter into the likelihood function.

Parameter	Africana Old		Africana New	
	Estimate	Std. Error	Estimate	Std. Error
$\log \sigma_{\Phi}$	0.077	0.02	0.17	0.02
$\phi$	0.927	0.01	0.92	0.01
$\log \sigma_N$	-0.039	0.05	-0.08	0.05
$\log \sigma_S$	-3.145	0.25	-2.76	0.24

Table 1: Parameter estimates

### SI 1.2: Implementation

The statistical model in the previous defines the joint distribution of the count data,  $N$ , and the unobserved random variables  $\Phi$ ,  $R$ ,  $S$ , for given parameters  $\sigma_S$ ,  $\sigma_{\Phi}$ , and the two parameters (scale and range) defining the nugget effect. The unobserved  $\Phi$ ,  $R$  and  $S$  are integrated out using the Laplace approximation, to yield the likelihood function as a function of the four parameters. The likelihood function is maximized to yield estimates of the four parameters, after which the posterior means of the  $\Phi$ ,  $R$ , and in particular  $S$  are reported.

The computations are performed in R version 3.1.2 (R Core Team, 2015); we use the Template Model Builder (TMB) package (Kristensen et al., 2016) for evaluating the likelihood function and its derivatives.

### SI 1.3: Data

The data base consisted of a total of 236 pairs of trawl hauls performed by RVs Africana and Dr. Fridtjof Nansen. The Gisund gear was used onboard Fridtjof Nansen, while RVs Africana deployed two gear types: ‘‘Africana Old’’ (108 hauls) and ‘‘Africana New’’ (128 hauls). Catch in numbers per length group and the swept area (hauling distance multiplied by wing spread) were available for each haul.

### SI 1.4: Results

The obtained intercalibration curves are seen in figure 1. Notice that Gisund overall is more effective than both the Old and the New Africana, in particular in the small size classes. The difference between size classes is statistically significant ( $p < 10^{-4}$ ). For size classes larger than 30 cm, say, the intercalibration curves show little variation with size and although this has not been tested, it is plausible that these variations are not statistically significant. Estimated parameters are seen in table 1.

## References

Kristensen K, Nielsen A, Berg C, Skaug H (2016) TMB: Automatic differentiation and Laplace approximation. *J. Stat. Softw* 70.



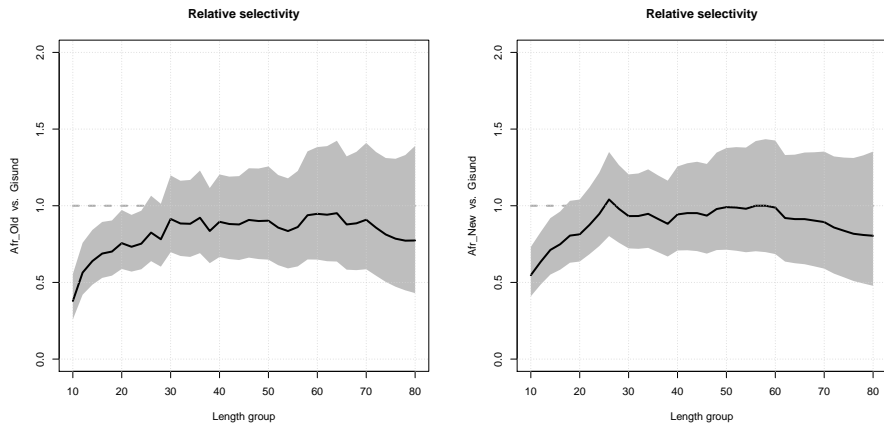


Figure 1: Relative selectivity (gear calibration factor), comparing catches of *M. paradoxus* with Gisund gear and the “Old” and “New” gear on the R/V Africana. Large values indicate that Africana is more effective. Solid curve: Estimated relative selectivity (posterior mode). Grey region: Marginal 95 % confidence intervals.

R Core Team (2015) *R: A Language and Environment for Statistical Computing* R Foundation for Statistical Computing, Vienna, Austria.

## Supplementary information 2

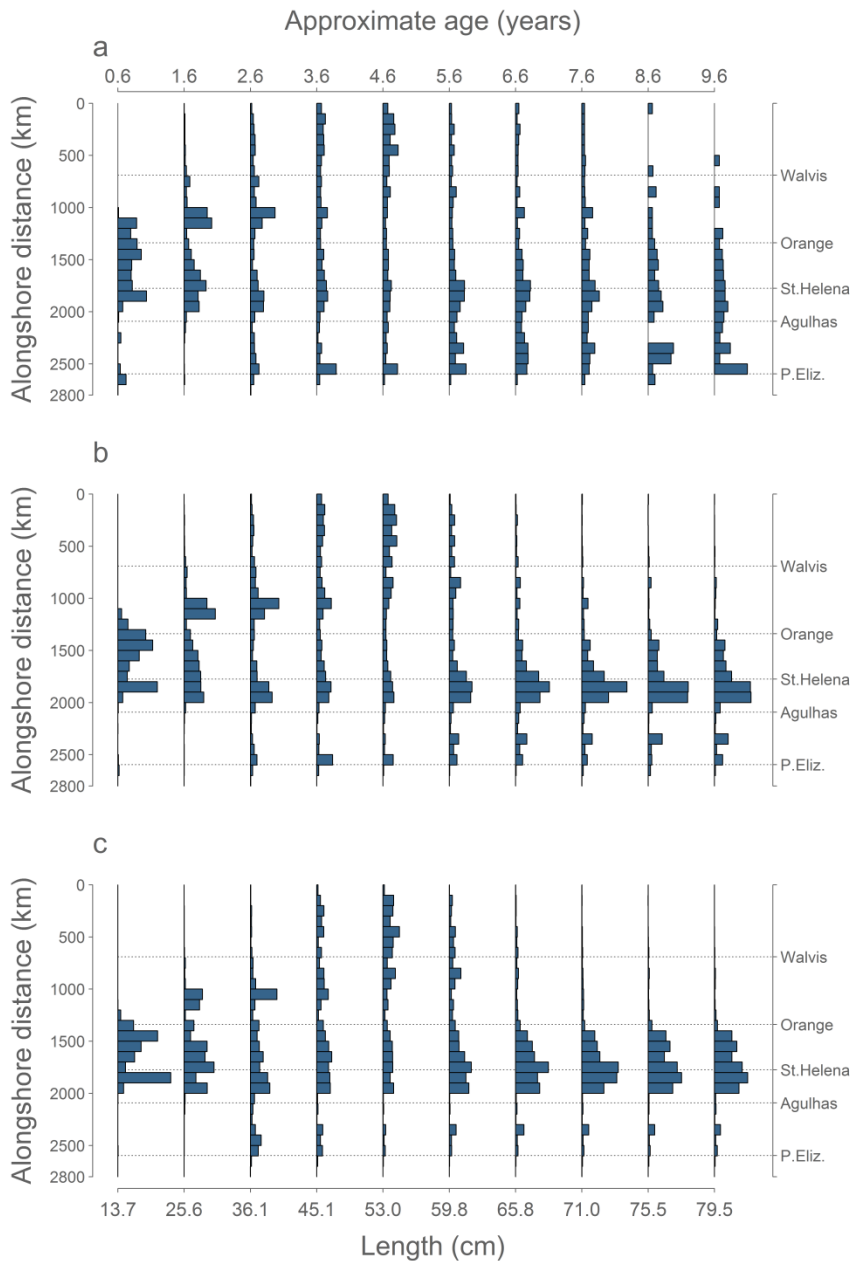


Figure SI2.1. Alongshore distribution of number of deep-water hake (*M. paradoxus*) by size and approximate age. Simple average of catches (a), GeoPop model fit (b) and GeoPop model after removal of local noise (nugget effect) (c). The spatial distributions have been projected onto a curvilinear axis following the coastline from the Namibia-Angola border (Kunene River) in the North to Port Elizabeth in the south-east. This figure is made for comparison with figure 5 in Jansen et al. (2016) and is therefore made in the same design (Jansen et al., 2016).