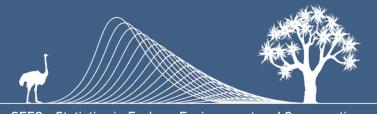
Distance sampling: Estimating densities of wildlife populations by modelling detectability



David Maphisa & Florian Weller





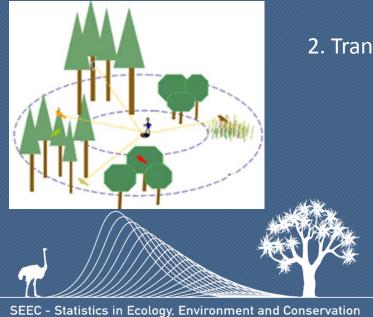




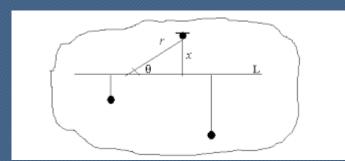
Distance sampling is applicable to many species groups (birds, plants, mammals, whales etc.) Surveys can be done on foot, plane, boat etc.

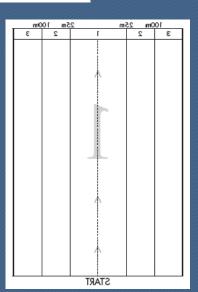


1. Point counts



2. Transects counts







The key is to account for observation process

Many factors can determine your ability to observe the species accurately eg. vegetation (grassland birds)

Time of the day (or temp)

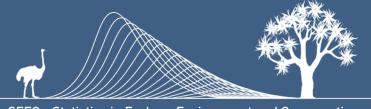
Cloud cover

etc.....

How many animals can you see on the 2nd pic below?









We present two approaches to analysing your distance data

1. Package unmarked (function disamp) in R (Chandler et al. 2015)

I demonstrate this by using bird and vegetation data from my Phd thesis

My study site (FS-KZN boundary)





2. Florian will demonstrate Program DISTANCE (standalone) and talk more about assumptions and biases





The way package unmarked is implemented in @



Allows inclusion of factors that may affect your ability to see or not see animals etc

Hierarchical distance sampling

include factors that affect density





Data requirements

- Bird distance sampling data (in my case fixed distance bands)
- 2. Habitat data in my case detctn & density are affected by
- (i) fire whether each transect was burned or not burned
- (ii) Grazing in 3 categories (none, light or heavily grazed)
- (iii) My main focus was on grass height (avh) and grass cover (cover)

all these datasets (birds and vegtn must be loaded in R) – merged into one object

Once this is done actual analysis is carried out (first detection and then density)





Out of eight species – work on one spcs at a time

distances

```
data PA <- bldata[bldata$Species=="African Pipit",]</pre>
data.PA$Season <- factor(data.PA$Season)
summary(data.PA)
ITUMF <- with(data.PA, {
unmarkedFrameDS(y = cbind(X50m, X100m, X100m.1),
siteCovs = data.frame(Year, Season, Area),
dist.breaks = c(0, 50, 100, 500)
tlength = rep(500,dim(data.PA)[1]), survey = "line", unitsln = "m")})
ItUMF
summary(ItUMF)
                                                  Histogram of distances
                                             0.015
hist(ItUMF)
                                             0.010
                                           Density
                                                           300
```

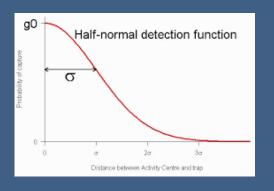


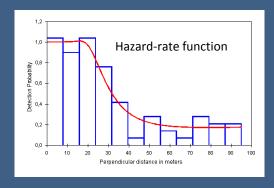
Model fitting – testing which model best fit your data Several functions are available

 $(LcCm1_default <- distsamp(~ 1 ~ 1, ltUMF))$

(LcCm1 halfNorm <- distsamp(~ 1 ~ 1, ltUMF, keyfun = "halfnorm", output = "density", unitsOut = "ha")) #

(LcCm1_hzml <- distsamp(~ 1 ~ 1, ltUMF, keyfun = "hazard", output = "density", unitsOut = "ha")) # Upon running the above codes /models - AIC values are produced to decide on best fit functn









For demonstration purposes this is what I did for African Pipit

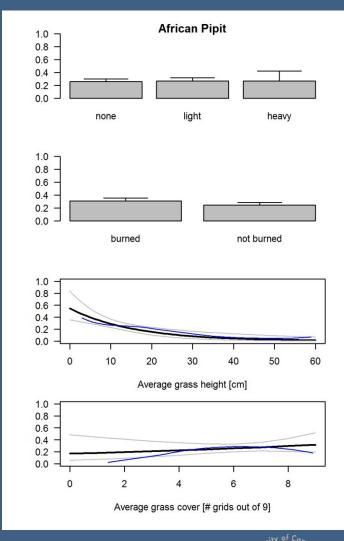
```
PAmdeten <- distsamp(~ (avh + cover) ~ 1, ItUMF))

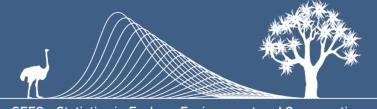
(PAm8 <- distsamp(~ (avh + cover) ~ Grazing, ItUMF))

(PAm9 <- distsamp(~ (avh + cover) ~ Burning))

(PAm10 <- distsamp(~ (avh + cover) ~ avh, ItUMF))

(PAm11 <- distsamp(~ (avh + cover) ~ cover, ItUMF))
```







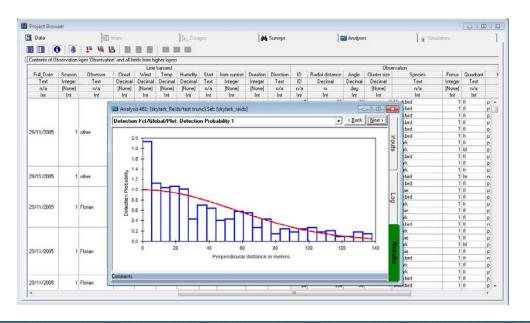
- Multiple covariate distance analysis with program DISTANCE
 - example 1: data in intervals
 - example 2: exact data
- More about the basic assumptions of distance sampling
 - how to deal with possible violations
- General considerations about the modeling process



Distance analysis with program DISTANCE

DISTANCE is a standalone Windows program for designing and analysing DS surveys. It is developed and maintained by a group based at the Centre for Research into Ecological and Environmental Modelling (CREEM) at the University of St Andrews.

- the same group maintains a number of DS R packages (primarily package *Distance*)
 - (note, this is a different implementation than unmarked)
- compatible with R on various levels; several of its components are in fact R modules
 - package readdst can port data and models directly from DISTANCE to R







Distance analysis with program DISTANCE

DISTANCE is a standalone Windows program for designing and analysing DS surveys. It is developed and maintained by a group based at the Centre for Research into Ecological and Environmental Modelling (CREEM) at the University of St Andrews.

DISTANCE vs R

Pro:

- easier entry and learning curve (learning R and DS at the same time can be daunting)
- better documentation (IMO)
- convenient output, analysis and error checking options
- survey design functionality (using spatial data)

Con:

- R allows greater customization and better data handling
- development happens primarily with R packages and is then ported, thus DISTANCE can be expected to lag a little in development







Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

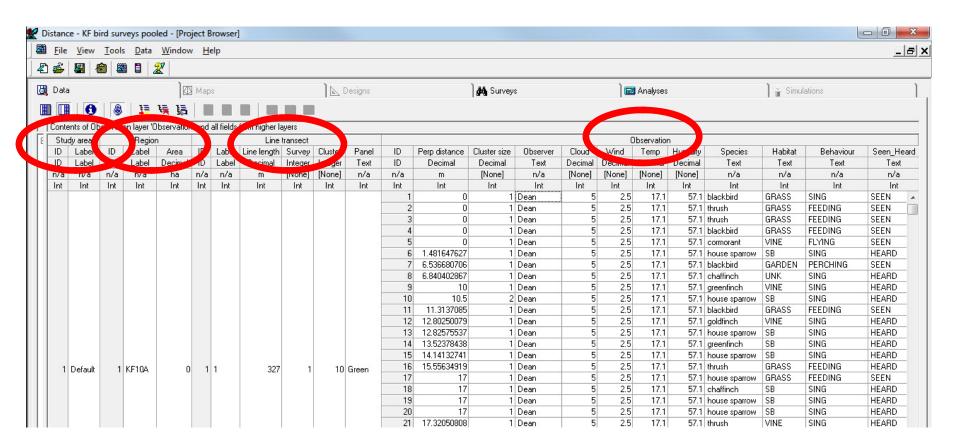
• import data

- store data in external file and import (using import wizard), rather than entering it directly in DISTANCE
- set up data filter and model definition
- run analyses as combinations of filters and definitions





Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand



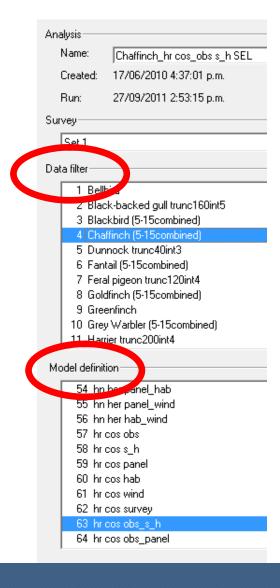






Chaffinch data from multi-species bird surveys in kiwifruit orchards, New

Zealand









Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Data selection Intervals Iruncation Units Distance intervals ✓ Transform distance data into intervals		sis
Number of intervals: Interval cutpoints Manual Automatic equal intervals	0 1 2 3 4	Cutpoints 0 15 25 50 100

The state of the s
Data selection Intervals Iruncation Units
Truncation of manually selected astarice intervals
Right truncation - choose from interval cutpoints C Right truncate at largest observed distance
C Discard the largest percent of distances
© Discard all observations beyond 100
Left truncation - choose from interval cutpoints
C No left truncation
© Discard all observations within 0
Truncation for cluster size estimation (where required)
Right truncation - choose from interval cutpoints
Same as that specified above
C Discard all observations beyond 100







Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

	MCD3 - Multiple covariates distance sampling	
donathan Jordan / Wikmedia Commons	Estimate Detection function Cluster size Multipliers Variance Misc.	
Analysis Engine: MCDS - Multiple covariates distance sampling Extincts Detection function Cluster size Multipliers Variance Misc. Models Adjustment terms Covariates Constraints Diagnostics Detection function Series expansion 1 Hazard-rate Cosine Analysis Engine: MCDS - Multiple covariates distance sampling Estimate Detection function Cluster size Multipliers Variance Misc. Models Adjustment erms Covariates Corptraints Diagnostics	Models Adjustment terms Covariates Constaints Diagnostics Chi-sq GOF tests and histograms of distances Intervals (These settings are ignored if intervals are specified in the Data Filter) Automatic selection of intervals and number of tests Manual selection Number of tests MCDS - Multiple covariates distance sampling Estimate Detection function Cluster size Multipliers Variance Misc. Interval cutpoin Manual Models MCDS - Multiple covariates distance sampling Estimate Detection function Cluster size Multipliers Variance Misc. No stratification No stratification Stratum Field name: Post-stratify, using: Stratum Sample definition (for encounter rate)	_
Detection function covariate Layer type containing	Output KS tests an Use layer type: Sample Create file of histog Quantities to estimate and level of resolution	
covariate Field name of covariate Factor Cluster si.	size File name: T:\Wo	
Observation Observer ✓ □	Level or resolution or estimates	
Observation Seen_Heard ✓	Global Stratum Sample	
	Density 🗸	
	Encounter rate	
	Detection function ☑ □ □ □	
	Cluster size (if required)	
Cluster size To include cluster size as a covariate, add the cluster size field to the table of covariates and tick the 'Cluster size' box in that row. When cluster size is a covariate, density is estimated using a different algorithm (see Help for details). Options are changed in the Estimate, Cluster Size and Variance tabs.	Global density estimate is Mean of stratum estimates weighted by Total effort in stratum Strata are replicates	





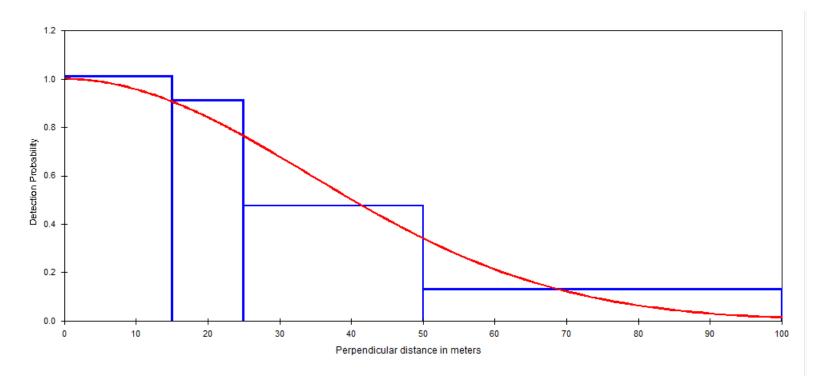


Example analysis:

Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Basic model (halfnormal function, no covariates)

• detections are in 4 distance bands

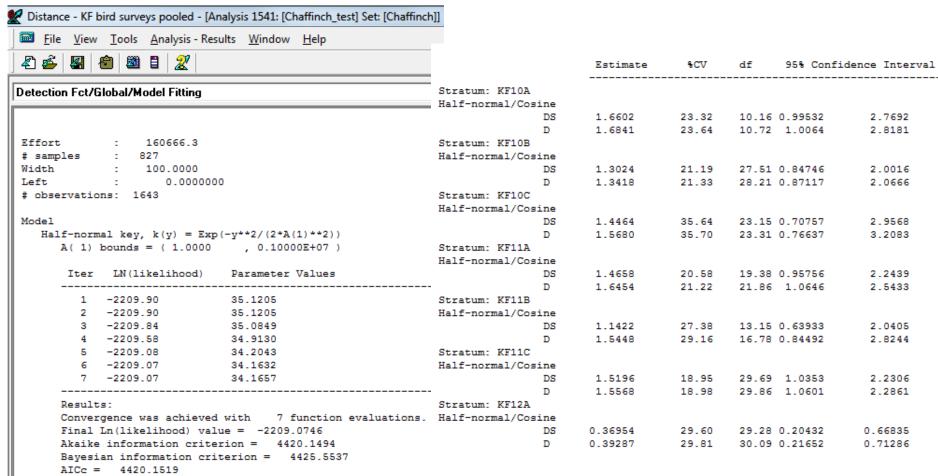








Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

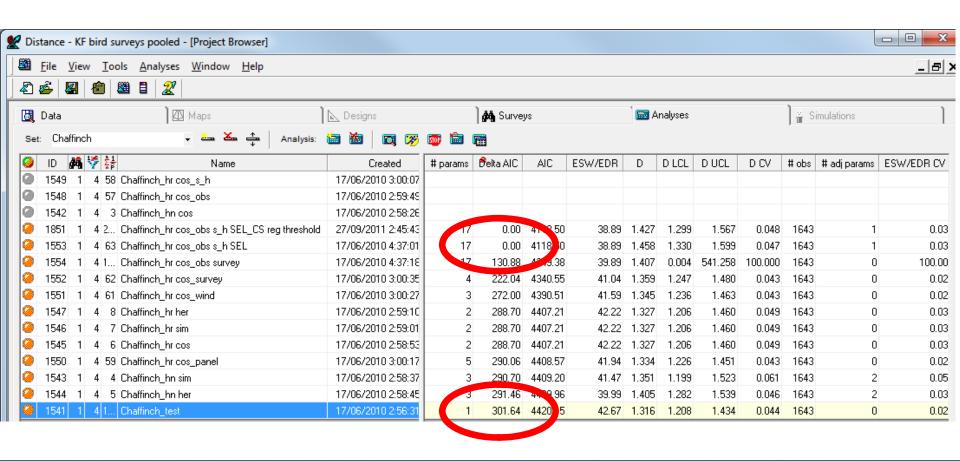








Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand





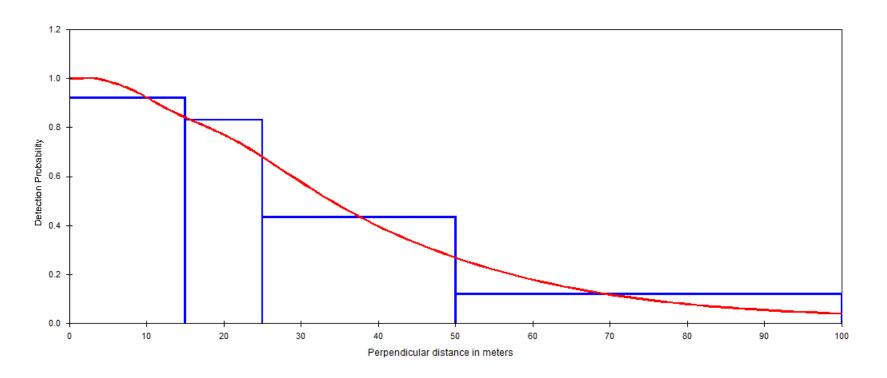




Example analysis:

Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Hazard rate function, covariate: observer (average of covariate levels)





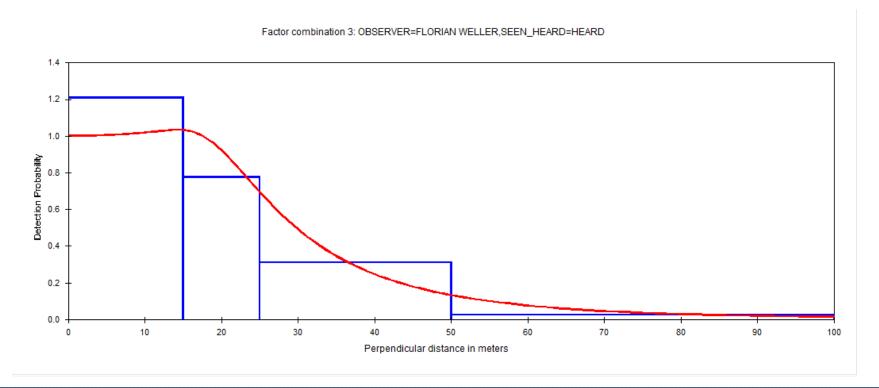




Example analysis:

Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Hazard rate function, covariate: observer (covariate level 1)



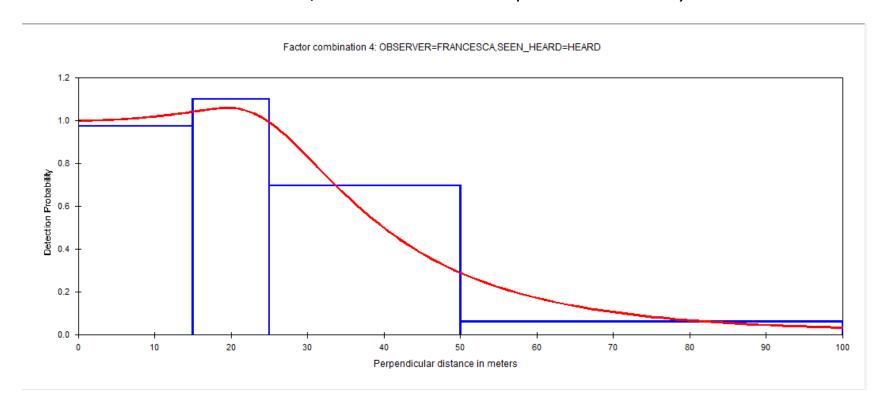






Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Hazard rate function, covariate: observer (covariate level 2)



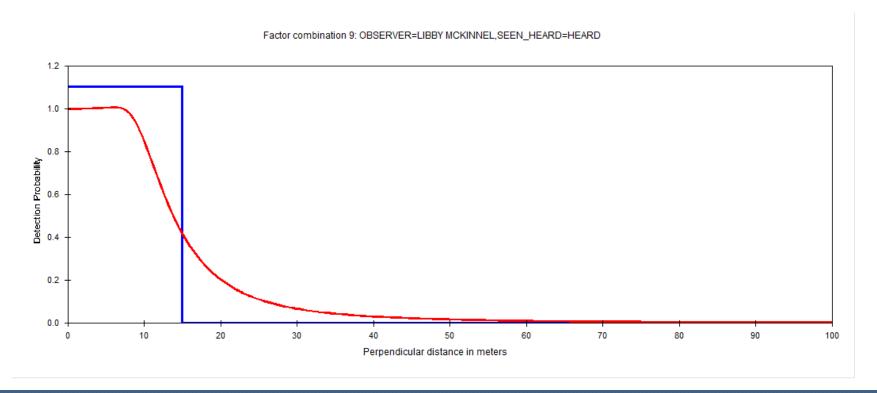






Chaffinch data from multi-species bird surveys in kiwifruit orchards, New Zealand

Hazard rate function, covariate: observer (covariate level 3)

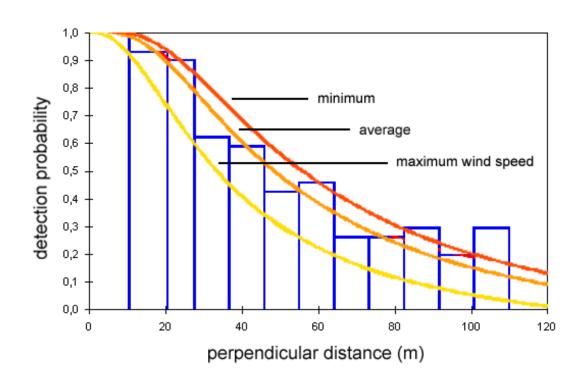








Covariates may be factorial (e.g. observer) or continuous (e.g. wind speed)









Covariates may be factorial (e.g. observer) or continuous (e.g. wind speed)

Keep in mind:

all covariates that enter into this model fitting process are covariates of detectability, not abundance. Here we are only concerned with effects on the detection process, to derive a largely unbiased survey result.

Many covariates can of course affect *both* detectability and abundance (e.g. site identity, vegetation type, time of year,...). However, modelling their effects on density/abundance itself is a separate (later) job, and not part of DS proper.

(unmarked combines these steps sequentially, but they are still different processes)

perpendicular distance (m)







Skylark data from multi-species bird surveys on sheep and beef farms, New Zealand

While the chaffinch data was collected in specific **distance bands**, the skylark detection distances (collected with a range finder) were analysed as **exact measurements**.

Pro:

- biases in detection process are easier to identify in high resolution data
- better model fit can be achieved

Con:

- requires a large number of detections to be feasible
- prone to overfitting
- you will likely end up partitioning the data into intervals anyway, to deal with biases (see below)
- → some partitioning into intervals / detection bands is generally a good idea; if not in the field then during the analysis stage





Allowing the detection function to fit more closely to the data

Detection function	Form
Uniform	1/w
Half-normal	$\exp(-y^2/2\sigma^2)$
Hazard-rate	$1 - \exp(-(y/\sigma)^{-b})$
Negative exponential	exp (- <i>ay</i>)

where y is distance and w is truncation distance

Extended form: detection function = key function + series expansion g(y) = key(y)[1 + series(y)]

Series expansions add **flexibility** to the shape of the key function

Cosine
$$\sum_{j=1}^{m} \alpha_{j} \cos(j\pi y/w)$$
 Simple polynomial
$$\sum_{j=1}^{m} \alpha_{j} (y/w)^{2j}$$
 Hermite polynomial
$$\sum_{j=2}^{m} \alpha_{j} H_{2j} (y/\sigma)$$

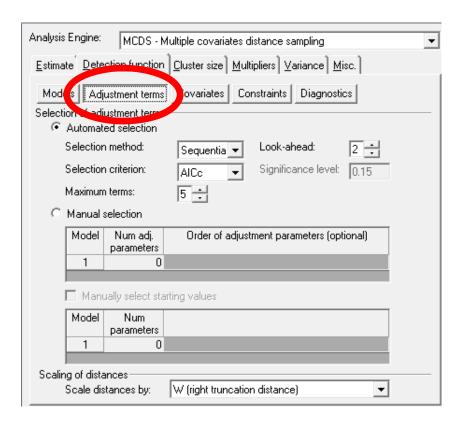
Simple polynomial
$$\sum_{i=1}^{m} \alpha_j (y/w)^2$$

Hermite polynomial
$$\sum_{j=2}^{m} \alpha_{j} H_{2j}(y/\sigma)$$





Allowing the detection function to fit more closely to the data

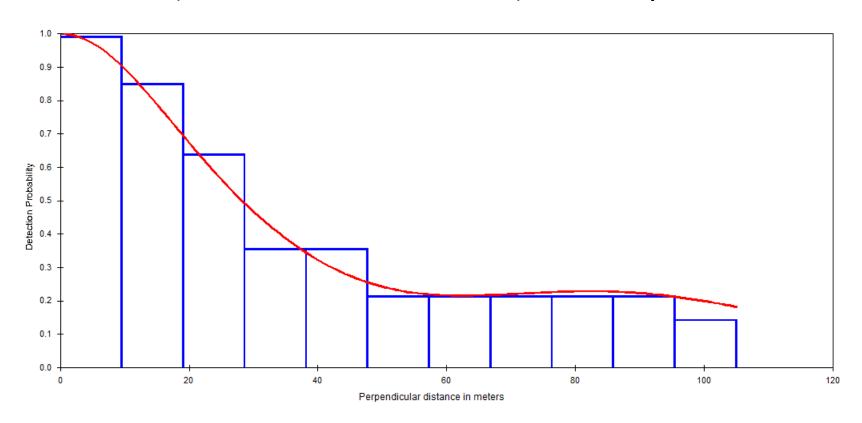






Song thrush data

Basic model (halfnormal function, no covariates) + 2 series expansion terms



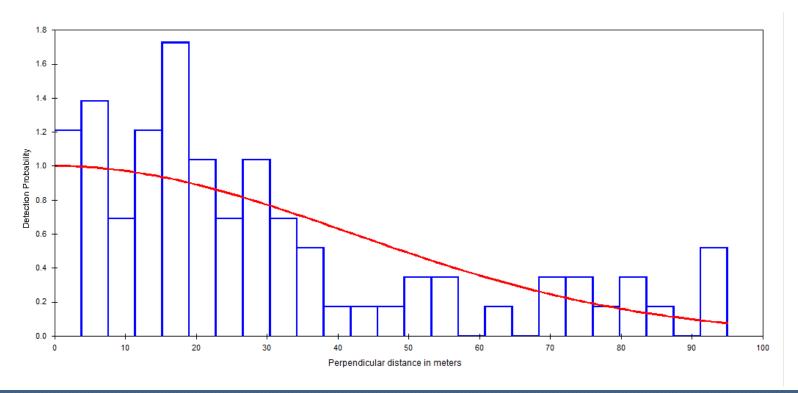






Skylark data from multi-species bird surveys on sheep and beef farms, New Zealand

Basic model (halfnormal function, no covariates)



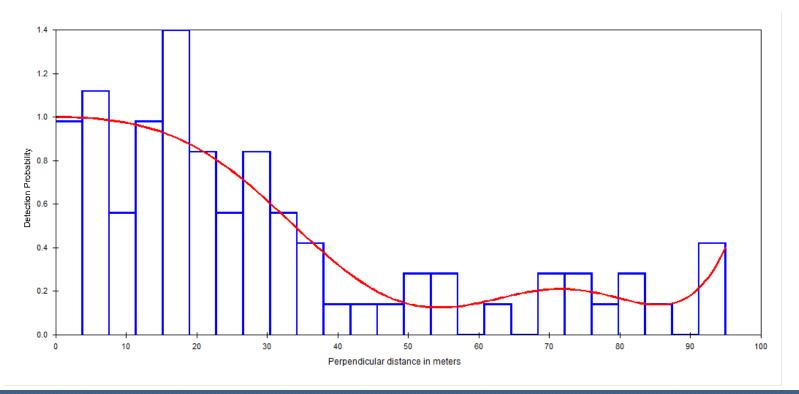






Skylark data from multi-species bird surveys on sheep and beef farms, New Zealand

Basic model (halfnormal function, no covariates) + 5 series expansion terms -> overfitting





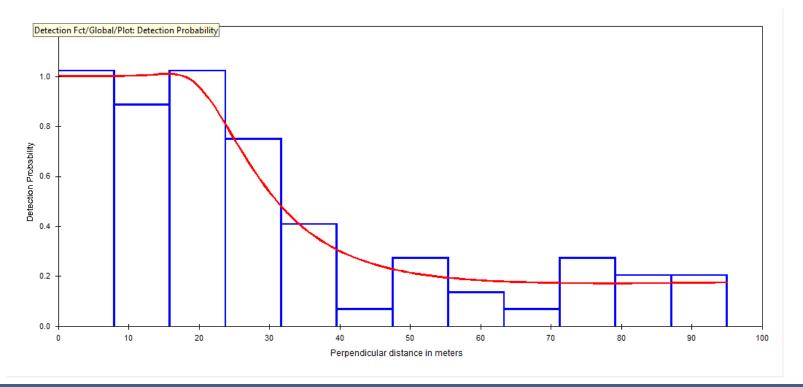




Skylark data from multi-species bird surveys on sheep and beef farms, New Zealand

Hazard rate function, no covariates

• partitioned into 8 m intervals







- Basic assumptions of distance sampling
- General considerations about model choice





Assumption #1:

All animals on the line are detected (i.e., detection probability at distance 0 is 1)

- depends on study subject; often an issue in shipboard surveys because ship may block field of view, and/or animals may dive
- in terrestrial surveys, can generally be assumed
- animals may **move away** from the line before observer passes that point; however, this does not result in a non-detection but in **movement before detection** (see below)

dealing with violations:

- this is tricky, because g(0) = 1 is a fundamental assumption of the fitting process
- can use a 2nd observer to "guard the centerline" (may result in **over**estimation)
- there are procedural fixes but they aren't pretty
- try **not** to violate this one





Assumption #2:

Animals are randomly and independently distributed

- three sources of bias:
 - a) populations are **clustered** (flocks etc.) but individual detections are treated as independent
 - artificially shrinks the confidence interval of the detection estimate
 - b) transects are not placed independently of **gradients of density** (e.g. roads)
 - can lead to strongly biased detection at specific distances
 - c) transects are too close together

dealing with violations:

- a) record not individuals but **clusters + cluster size**, then incorporate cluster size into the detection funtion
- b) place transects either randomly, or across gradients of density
- c) make sure that maximum detection ranges do not overlap between transect





Recording clusters instead of individuals

Effort : 34965 # samples : 71 Width : 80.00000 # observations: 157

Half-normal key, $k(y) = \text{Exp}(-y^{**2}/(2^*A(1)^{**2}))$ A(1) bounds = (0.80000 , 0.10000E+07)

Iter	LN(likelihood)	Parameter Values
1	-649.744	30.6225
2	-649.744	30.6225
3	-649.635	30.9778
4	-649.496	31.9478
5	-649.494	32.0888
6	-649.494	32.1064
7	-649.494	32.1067

Results:

Convergence was achieved with 7 function evaluations.

Final Ln(likelihood) value = -649.49381 Akaike information criterion = 1300.9877 Bayesian information criterion = 1304.0438

AICc = 1301.0134

Analysis Engine:	MCDS - M Tiple cover has distance sampling
Estimate Detect	ion fur tion Cluster size Multi liers Variance Misc.
Cluster size estima (These options	ation method s are ignored unless objects are clusters.)
Use size-b	ias regression method
C Use mean	of observed clusters
C Use size b alpha-level	ias regression method if regression significant at an of 0.15 ; use mean of observers if not significant
_ Size-bias reg	ression method
Regres	s In(cluster size) against estimated g(x)
C Regres	s cluster size against estimated g(x)
C Regres	s In(cluster size) against distance x
C Regres	s cluster size against distance x

	Estimate	%CV	df	95% Confidence Interval	
Average cluster siz	1.1911	4.62	156.00	1.0872	1.3049
r r-p E(S		2.55	155.00	1.0808	1.1952





Assumption #3:

Animals do not move before detection

- bias resulting from movement is **negligible if movement is random**
- movement in response to the observer can result in negative (avoiding observer) or positive (attracted to observer) bias in detectability
- particularly avoidance behaviour is quite common

dealing with violations:

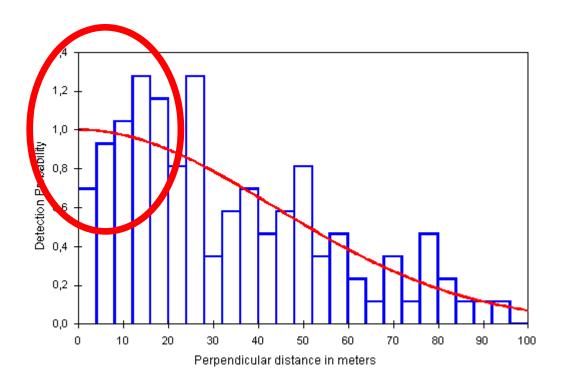
- average out & bridge dips/spikes in the distribution by
 - partitioning data into **intervals**
 - using models with a shoulder (e.g. hazard-rate)





Skylark data

Indications of observer avoidance in the data







Assumption #4:

Measurements are exact (angles and distances)

- the effect of **random** errors is negligible, but **systematic** errors can introduce bias; this often happens with **rounding**
- rounding angles or distances to preferred values can result in heaping
- angle rounding to zero is particularly common

dealing with violations:

- avoid dead reckoning in the field by using tools: range finders, angle boards, etc.
- smoothen & average instances of rounding in the data by partitioning into intervals

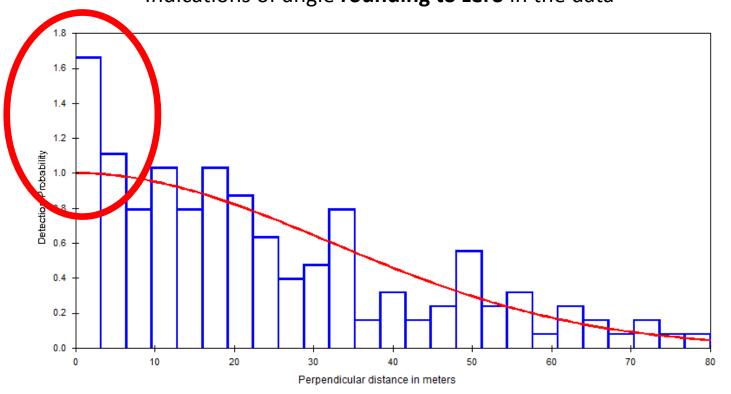






Blackbird data

Indications of angle rounding to zero in the data



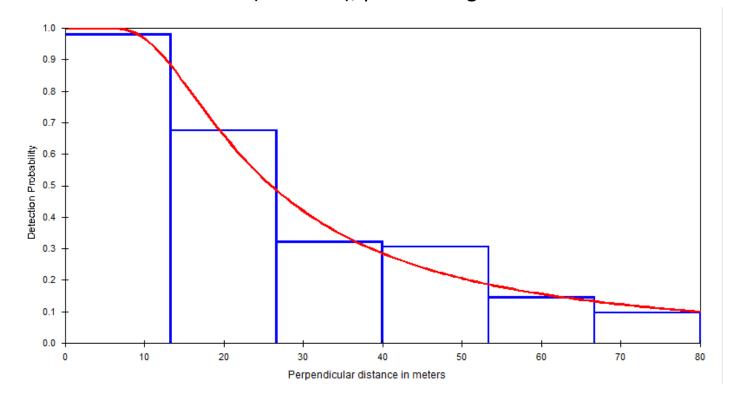






Blackbird data

Hazard rate (shoulder), partitioning into intervals







When to use distance sampling?

1

available number of samples / accuracy of abundance estimate

Census

(complete count)

Double count

(rapid large-scale counts, calibrated by a few small-scale censuses)

Distance sampling

(modelled detectability)

- + large improvement in accuracy by modelling detectability
- + much lower effort than census
- requires reasonably high number of records (>60)

Index count

(count not corrected for detectability, may be sufficient for tracking relative changes over time)

Presence/absence

(e.g. Occupancy)





When to use *multiple covariate* distance sampling? (as opposed to basic DS)

-> this is a question of level of detection function vs level of abundance/density estimate

Assume e.g. that one species was surveyed at several sites, and you want an estimate for each site. There are three basic options:

- a) Fit a separate detection function to each site dataset
 - → most accurate outcome, best adaptation to different site parameters; <u>BUT</u> requires sufficient sample sizes at **each** site
- **b)** Fit a **global** detection function using all pooled detection, then use this function on **site** datasets to get site-level estimates
 - → may save your bacon if sample sizes at individual sites don't support fitting a function; <u>BUT</u> assumes that detection does **not vary** (substantially) between sites
- **c)** Use **multiple covariate distance sampling**: fit a global detection function, specifying site as a **covariate** (i.e., allow the global function to vary for data subsets defined by covariate levels)
 - → this is halfway between the other two options in requirements and result accuracy

The same choice applies to any covariate (canopy cover, season, sex, etc.)

→ this is essentially a **model selection question**; if the choice isn't clear, AIC can be used to help choose between approaches (as all three use the same data set)





Useful resources

• The Distance project website

- http://distancesampling.org/
- software downloads (R packages & program Distance)
- Distance mailing list / Google group
- huge distance sampling article bibliography
- Books (by Steve Buckland *et al.*):
 - Introduction to Distance Sampling: Estimating Abundance of Biological Populations. (2001) Oxford University Press. (in Main Library)
 - Advanced Distance Sampling. (2004) Oxford University Press.
 - Distance Sampling: Methods and Applications. (2015) Springer.
- Program Distance documentation (very useful)
- <u>Method reference</u>: Thomas, L., S.T. Buckland, E.A. Rexstad, J. L. Laake, S. Strindberg, S. L. Hedley, J. R.B. Bishop, T. A. Marques, and K. P. Burnham. 2010. Distance software: design and analysis of distance sampling surveys for estimating population size. Journal of Applied Ecology 47: 5-14. DOI: 10.1111/j.1365-2664.2009.01737.x



